

Can we achieve atmospheric chemical environments in the laboratory? An integrated model-measurement approach to chamber SOA studies

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Abstract

Secondary organic aerosol (SOA), atmospheric particulate matter formed from low-volatility products of volatile organic compound (VOC) oxidation, impacts both air quality and climate. Current 3D models, however, cannot reproduce the observed variability in atmospheric organic aerosol. Because many SOA model descriptions are derived from environmental chamber experiments, our ability to represent atmospheric conditions in chambers directly impacts our ability to assess the air quality and climate impacts of SOA. Here, we develop a new approach that leverages global modeling and detailed mechanisms to design chamber experiments that mimic the atmospheric chemistry of organic peroxy radicals (RO₂), a key intermediate in VOC oxidation. Drawing on decades of laboratory experiments, we develop a framework for quantitatively describing RO₂ chemistry and show that no previous experimental approaches to studying SOA formation have accessed the relevant atmospheric RO₂ fate distribution. We show proof-of-concept experiments that demonstrate how SOA experiments can access a range of atmospheric chemical environments and propose several directions for future studies.

Introduction

Atmospheric aerosols are integral to two of today's most important environmental concerns: air pollution and climate. Exposure to aerosol pollution is associated with over 8 million premature deaths each year, making it the leading environmental risk factor for premature mortality (1, 2). Atmospheric aerosols also impact the global climate by absorbing and scattering sunlight, as well as by altering cloud properties, with these impacts representing the largest source of uncertainty in our understanding of global radiative forcing (3). Organic aerosol (OA) constitutes a large, and sometimes dominant, fraction of fine aerosol mass (4). Much of this OA is secondary (secondary organic aerosol, SOA) (4, 5), produced from volatile organic compounds (VOCs) that are oxidized in the atmosphere to form lower-volatility species. Despite the importance of SOA in the atmosphere and the decades of SOA-focused laboratory experiments, field measurements, and modeling studies, the complex chemical processing that leads to SOA formation is still not well constrained (6).

The first laboratory experiments examining SOA formation were recorded by Tyndall in the late 19th century; he observed the formation of a 'blue cloud' when irradiating organic vapors in a glass tube and attributed his observations to the formation of particles (7). Haagen-Smit's 'smog chamber' experiments in the 1950s demonstrated that a mixture of vehicular hydrocarbon pollutants (e.g., alkenes), O₃, and NO₂ formed particles upon irradiation by sunlight, helping to explain the

51 chemistry of urban smog (8). Later, Went (9) extended these experiments to include the oxidation
52 of natural hydrocarbons, and postulated that photochemistry involving biogenic VOCs contributed
53 ‘blue hazes’ in the atmosphere. ‘Smog chamber’ experiments throughout the 1980s continued to
54 show evidence for SOA production from photochemistry involving biogenic and anthropogenic
55 VOCs with O₃ and NO_x (10, 11).

56
57 By 1990, researchers began to quantitatively determine SOA yields for oxidation of individual
58 VOCs (11, 12), which would eventually enable yield-based parameterizations in atmospheric
59 models. A dependence of SOA yields on NO_x concentrations emerged from these chamber
60 experiments, suggesting that atmospheric SOA formation depends on the chemical environment
61 (13–16). Because of this observed NO_x-dependence, studies began to characterize experiments by
62 their initial hydrocarbon-to-NO_x ratio as is common in descriptions of tropospheric ozone formation
63 (13, 14, 16). The mechanism for the dependence of SOA yield on this ratio was initially uncertain,
64 but by 1999 it was generally assumed to solely result from the changes in the relative importance
65 of different oxidants (e.g. OH, O₃, NO₃) with varying NO_x concentrations (17, 18). The first
66 regional and global models of SOA also reflected this understanding: NO_x influences on O₃ and
67 OH were included, but SOA yields for a given oxidant were assumed not to vary with NO_x
68 concentration (19–21).

69
70 In the early- and mid-2000s, chamber studies demonstrated that changes in SOA yields with NO_x
71 were a result of both changes in oxidant ratios and changes in the fate of RO₂, organic peroxy
72 radicals produced in most VOC oxidation processes (22). For example, chamber experiments
73 demonstrated that the relative importance of hydroperoxides (products of the reaction of RO₂
74 radicals with HO₂) to the oxidation product distribution could help explain the dependence of the
75 SOA yield on NO_x (23–25). With the understanding that RO₂ chemistry is influenced by NO_x
76 concentrations (26, 27), chamber experiments were designed to access ‘limiting conditions’ in
77 which the fate of RO₂ was dominated entirely by reaction with a single co-reactant: ‘low-NO’
78 experiments were designed to mimic ‘clean’ atmospheric conditions with RO₂+HO₂ as the
79 dominant RO₂ fate, whereas ‘high-NO’ experiments were designed to mimic ‘polluted’ atmospheric
80 conditions with RO₂+NO dominating RO₂ reactivity (27–32). The development of a chemical
81 coordinate to track the competition between NO and HO₂ for reaction with RO₂ followed; this
82 coordinate was gradually refined and by 2010 was defined by Pye et al. (33) as:

$$\beta = \frac{k_{RO_2+NO}[NO]}{k_{RO_2+NO}[NO]+k_{RO_2+HO_2}[HO_2]} \quad \text{Equation (1)}$$

84 Using SOA yields measured under limiting RO₂ fate conditions, β could be interpreted as a ‘mixing
85 parameter’ to estimate SOA yields at intermediate-NO_x conditions (34, 35). By the late 2000s, SOA
86 parameterizations in global models began to reflect the newly understood importance of NO_x for
87 RO₂ fate by incorporating β -based SOA parameterizations (33, 36–38).

88
89 Recent chamber experiments, however, have demonstrated that SOA yields may also depend on
90 three additional RO₂ fates: RO₂ isomerization, RO₂+NO₂, and RO₂+RO₂ reactions (Figure 1).
91 Laboratory and computational studies over the last decade have demonstrated that for some RO₂
92 species, unimolecular isomerization pathways have sufficiently short lifetimes (τ_{uni}) that they can
93 be competitive with bimolecular reactions (i.e., $\tau_{uni} \leq \tau_{bi}$, where τ_{bi} is the RO₂ lifetime to
94 bimolecular reactions) (39, 40). Such isomerization reactions and subsequent rapid molecular
95 oxygen additions can lead to the production of highly oxidized products, which can contribute to
96 the formation of low-volatility SOA (41). Recent chamber studies have also demonstrated that the
97 NO/NO₂ ratio during an experiment, which controls what fraction of acyl-RO₂ (RC(=O)O₂) react
98 with NO versus NO₂, has a measurable effect on the oxidative product distribution and/or amount
99 of SOA formed. At lower NO/NO₂ ratios, RO₂+NO₂ reactions can dominate the chemistry of acyl-

100 RO₂ radicals, thereby altering the oxidative product distribution and increasing the formation of
101 peroxy acyl nitrates (PANs) (42, 43) which can have an influence on SOA formation (44). Other
102 chamber studies have demonstrated that SOA yields are affected by the RO₂/HO₂ ratio, which
103 controls what fraction of RO₂ react with HO₂ versus with other RO₂ radicals (23, 45–47). The
104 direction and magnitude of this effect on SOA yields is determined by the volatility of RO₂+RO₂
105 products, which include alkoxy radicals (RO+RO), alcohol and carbonyl products (ROH+R'CHO),
106 and organic peroxides (ROOR)). Numerous studies have highlighted ROOR products in particular
107 as potential contributors to SOA formation (48–50). As such, current evidence indicates that four
108 bimolecular RO₂ fates (reactions with NO, HO₂, NO₂, and RO₂), as well as unimolecular RO₂
109 reactions, may all affect SOA formation.

110
111 Current model parameterizations of SOA that include NO_x dependence, however, are still based on
112 chamber experiments run under high-NO and low-NO extremes, with the untested assumption that
113 SOA yields can be parameterized as a linear combination of yields measured at the two limiting
114 cases ($\beta=0$ and $\beta=1$). By design, these ‘limiting condition’ experiments include high radical
115 abundances that shorten the RO₂ bimolecular lifetime, thereby limiting RO₂ isomerization, and do
116 not control for the role of RO₂+NO₂ or RO₂+RO₂ reactions. Moreover, experiments run at limiting
117 ‘high-NO’ and ‘low-NO’ conditions that underlie today’s model parameterizations are likely not
118 fully relevant for SOA produced over multiple generations in the atmosphere, since the RO₂ fate
119 might differ from one generation of oxidation to the next. As a result, current model
120 parameterizations of SOA are mimicking SOA produced under idealized chamber conditions rather
121 than representing realistic atmospheric chemical environments. This likely contributes to the
122 inability of current large-scale models to reproduce the observed variability in atmospheric OA (51,
123 52).

124
125 Because model parameterizations are informed by laboratory experiments, model accuracy hinges
126 on the atmospheric relevance of both the physical and chemical conditions of laboratory
127 experiments. Porter et al. (53) identified that chamber studies of SOA are often run under dry, room-
128 temperature conditions, which are not representative of much of the atmosphere, and began to
129 quantify what chemical space chamber experiments of SOA need to match in order to represent the
130 atmosphere. Further quantification of atmospheric and chamber chemical environments and
131 development of strategies for maximizing overlap between the two is the focus of this work.

132
133 Given the importance of bimolecular and unimolecular RO₂ chemistry for SOA production, here
134 we define a four-parameter chemical space that allows us to track the fate of RO₂ in both the
135 atmosphere and in laboratory studies. We use these parameters in conjunction with global modeling
136 to define the chemical environments in which SOA is formed in the atmosphere, and we use
137 mechanistic box modeling of environmental chambers to assess what parts of this chemical space
138 are accessible in the laboratory. We discuss the challenges associated with mimicking the
139 atmospheric chemical environment in chamber studies, show proof-of-concept experiments that
140 demonstrate how SOA experiments can access a range of atmospheric chemical environments, and
141 present pathways forward for furthering our understanding of SOA production in the atmosphere
142 in light of the challenges.

143 **Results**

144 *Atmospheric RO₂ fate distribution*

145
146
147 A complete, quantitative description of RO₂ fates is needed to fully characterize an atmospheric
148 chemical environment (54). Though β describes the competition between RO₂+HO₂ and RO₂+NO
149 reactions, the contributions of RO₂ isomerization, RO₂+RO₂ reactions, and RO₂+NO₂ reactions to

150 atmospheric RO₂ reactivity necessitate additional chemical coordinates to fully describe the RO₂
151 fate distribution. Here we combine four parameters (β , τ_{bi} , RO₂/HO₂, and NO/NO₂) which have
152 been used independently to characterize subsets of RO₂ reaction pathways (23, 33, 44, 53, 55, 56)
153 to complete a parameter space that fully describes atmospheric RO₂ fates as outlined in Figure 1.
154 To our knowledge, this is the first time that these parameters have been considered together, and
155 that more than one has been considered at a time, to describe RO₂ chemistry in the atmosphere and
156 within chambers.

157
158 We adopt the bimolecular RO₂ lifetime against reaction with NO and HO₂, as originally defined by
159 Teng et al. (2017):

$$\tau_{bi} = \frac{1}{k_{RO_2+NO}[NO] + k_{RO_2+HO_2}[HO_2]} \quad \text{Equation (2)}$$

161 A comparison between τ_{bi} and τ_{uni} , the RO₂ lifetime to unimolecular isomerization, indicates the
162 relative contribution of bimolecular and unimolecular reactions. We also adopt the RO₂/HO₂ ratio
163 to quantify the relative importance of RO₂+RO₂ chemistry and thus ROOR formation (23), as well
164 as the NO/NO₂ ratio to denote the relative importance of PANs to the RO₂ fate (43, 44). Ideally a
165 ratio of rates (analogous to β) instead of a ratio of radical concentrations (RO₂/HO₂ and NO/NO₂)
166 would be used to define the relative importance of RO₂+RO₂ and RO₂+NO₂ reactions to the RO₂
167 fate distribution, and a complete τ_{bi} definition would also include the lifetime of RO₂ against
168 reaction with other RO₂ and with NO₂. However, because RO₂+RO₂ rate constants are both highly
169 variable and uncertain (48, 49), here we use the RO₂/HO₂ ratio to indicate the relative importance
170 of RO₂+RO₂ reactions and omit RO₂+RO₂ reactions from the τ_{bi} determination (Eq. 2). Similarly,
171 because RO₂+NO₂ reactions only produce stable products for acyl-RO₂, we use the NO/NO₂ ratio
172 instead of a ratio of rates to simplify accounting and omit RO₂+NO₂ reactions from the τ_{bi}
173 calculation, noting that τ_{bi} will be shorter for acyl-RO₂ than for other RO₂ (Eq. 2).

174
175 To examine the fate of RO₂ in the atmosphere, we focus our analysis here on the oxidation of
176 isoprene. Multi-generation isoprene oxidation is an important contributor to global SOA production
177 (36), and the gas-phase oxidation chemistry of isoprene (including its RO₂ chemistry) has been
178 studied extensively over the last decade (57). We also include a parallel set of analyses for
179 monoterpene-derived SOA (Figures S1-S2); the results for monoterpenes are similar to those
180 presented here for isoprene.

181
182 Understanding atmospheric RO₂ chemistry requires identifying which parts of the 4D chemical
183 space defined here are populated during VOC oxidation in the atmosphere. Figure 2 shows the
184 atmospheric distribution of β , τ_{bi} , RO₂/HO₂, and NO/NO₂ for isoprene-derived RO₂, calculated from
185 hourly output from the GEOS-Chem chemical transport model (version 13.4.0,
186 <https://doi.org/10.5281/zenodo.7254268>) for January and July 2016. We show the entire calculated
187 global distribution of these four parameters, weighted by the rate of isoprene+OH oxidation; as
188 such, any references to the ‘atmospheric distribution’ refer to the atmospheric distribution in regions
189 where isoprene is undergoing OH-initiated oxidation (and any references to isoprene oxidation refer
190 to OH-initiated oxidation), unless otherwise specified.

191
192 We note that the use of a global chemical transport model means our analysis is inherently at coarse
193 spatial resolution (here 2°x2.5°), a scale relevant for considering global SOA production where
194 most isoprene oxidation occurs. However, the model does not capture the distribution of RO₂ fates
195 at fine spatial scales. For example, highly polluted environments would likely tend toward higher
196 β and lower τ_{bi} than seen in the global distribution, resulting in conditions where RO₂+NO reactions
197 are more likely to dominate the RO₂ reactivity. Even so, as NO_x emissions decrease in cities as a

198 result of emissions controls, values of τ_{bi} in urban areas are increasing (58), meaning future urban
199 chemical environments may be better reflected in the global distributions shown here.

200
201 As shown in Figure 2, β spans the full possible range but, in most regions of the global atmosphere,
202 does not lie at either extreme (median = 0.44, 25th percentile = 0.27, 75th percentile = 0.61). Rather,
203 the ‘intermediate- β ’ values that dominate the distribution indicate that RO₂ bimolecular reactions
204 with NO and HO₂ are in competition with each other throughout most of the global atmosphere.

205
206 Moreover, values of τ_{bi} mostly lie between 20 and 300 seconds throughout the atmosphere (Figure
207 2a). Within this distribution, τ_{bi} is generally longer in regions with lower β (where reaction with
208 HO₂ dominates) and shorter in higher- β environments (where reaction with NO dominates). The
209 relative importance of RO₂ isomerization is determined by the relative rates of each process (i.e.,
210 τ_{bi} vs. τ_{uni}). However, values of τ_{uni} are highly variable, as RO₂ isomerization rates are highly
211 structure-dependent and vary over orders of magnitude for different RO₂. In some cases,
212 isomerization is fast ($\tau_{uni} \ll 1$ s), meaning it will dominate throughout virtually the entire
213 atmosphere (55, 59). In contrast, the atmospheric fate of any RO₂ with very slow isomerization
214 rates ($\tau_{uni} > 1,000$ s) will be dominated by bimolecular reactions. RO₂ radicals with intermediate
215 τ_{uni} that are within an order of magnitude of atmospheric τ_{bi} (i.e., $1 \text{ s} < \tau_{uni} < 1,000 \text{ s}$), per contra,
216 may react by both bimolecular and unimolecular pathways in the atmosphere. Additionally, we note
217 that the temperature dependence of unimolecular isomerization rates is much steeper than the
218 temperature dependence of most bimolecular RO₂ reactions (39, 55); as such, the relative
219 importance of unimolecular and bimolecular reactions for a given RO₂ in the atmosphere can
220 change with temperature. However, the temperature variability for regions of the atmosphere with
221 substantial isoprene oxidation is relatively narrow (10th percentile = 285 K, median = 295 K, 90th
222 percentile = 302 K).

223
224 Figure 2b also shows that atmospheric RO₂/HO₂ ratios are largely below 1 and generally increase
225 as β decreases. Determining the relative importance of RO₂+HO₂ and RO₂+RO₂ reactions relies
226 critically on uncertain and variable RO₂+RO₂ rate constants. Assuming an RO₂+RO₂ rate constant
227 of 10⁻¹² at 298 K (49), an average atmospheric RO₂/HO₂ ratio of ~0.5 indicates that 5% of
228 RO₂+peroxy radical reactions in the atmosphere are RO₂+RO₂. Faster RO₂+RO₂ rate constants of
229 10⁻¹¹ and 10⁻¹⁰ (48) would suggest that 33% and 83%, respectively, of RO₂+peroxy radical reactions
230 are RO₂+RO₂.

231
232 The atmospheric NO/NO₂ ratio is centered around 0.4 (Figure 2c) and is largely controlled by O₃
233 abundance and NO₂ photolysis rates. In the GEOS-Chem mechanism, this average value
234 corresponds to net formation (production minus loss) of PANs in ~90% of atmospheric acyl-
235 RO₂+NO_x reactions at atmospheric temperatures.

236 237 *RO₂ fate distributions during SOA chamber experiments: previous experiments and challenges*

238
239 The global distributions of RO₂ fates in the atmosphere shown in Figure 2 can be viewed as
240 providing ‘targets’ for RO₂ chemistry in laboratory studies. However, as shown in Figure 3,
241 previous SOA chamber experiments have generally simulated RO₂ reactivities that differ
242 substantially from those in the atmosphere. The classic smog photochemistry chamber experiments
243 of the 1990s (e.g., Griffin et al., 1999), involving the irradiation of hydrocarbon-NO_x mixtures,
244 achieved the high- β conditions typical of highly polluted urban areas, but at much lower values of
245 τ_{bi} and NO/NO₂ than are typical in the atmosphere. The ‘limiting condition’ SOA experiments of
246 the mid-2000s (e.g., Kroll et al., 2005, 2006; Ng et al., 2007a) achieved the extreme- β conditions
247 they were designed for, but they did not capture all possible atmospheric RO₂ isomerization

248 products, nor did they fully match atmospheric distributions of RO₂/HO₂ and NO/NO₂ ratios (60,
249 61). Traditional dark ozonolysis experiments (62) did access atmospheric τ_{bi} , but achieved
250 RO₂/HO₂ ratios much higher than those found in the atmosphere. Recent ultraclean experiments
251 measuring HOM yields from ozonolysis at low NO concentrations in the CLOUD chamber (63)
252 accessed a range of β at longer RO₂ lifetimes, but also had RO₂/HO₂ ratios that far exceeded than
253 those of the atmosphere. Oxidation flow reactors (OFRs), which can be considered as a limiting
254 case for small chambers, have become increasingly popular for studying SOA formation over
255 longer aging timescales. Though OFRs can access the full range of β (64, 65), most OFR
256 experiments have accessed τ_{bi} for RO₂ that are shorter than those in the atmosphere (66, 67). Further
257 details about the chemical environments achieved in these previous approaches are included in
258 Section S1.

259
260 Though the aforementioned previous approaches to SOA studies span a wide range of RO₂ fates,
261 no previous approach has overlapped with the atmospheric distribution of RO₂ reactivity. The
262 longstanding challenges in matching atmospheric conditions in SOA chamber experiments derive
263 from two fundamental laboratory constraints (Figure 4): (1) a sufficiently high quantity of products
264 is needed to surpass instrumental limits of detection (LOD), and (2) oxidation timescales must be
265 fast enough to outcompete lab-specific loss processes, namely wall loss and dilution. In order to
266 produce enough OA to surpass instrument LODs ($\sim 1 \mu\text{g m}^{-3}$), experiments must start with large
267 quantities of the VOC precursor (typically 100s of ppbC). Together, the need for large precursor
268 concentrations and the requirement for fast oxidation timescales mean that experiments must be run
269 at relatively high oxidant levels, resulting in the production of high HO₂ and RO₂ concentrations
270 and/or requiring the addition of large amounts of NO. Ultimately, the chemical conditions of the
271 chamber are driven by the VOC oxidation itself, in contrast to the atmosphere where the ambient
272 chemical environment controls VOC oxidation conditions. When VOC oxidation controls the
273 chemical environment, which is the case not only in standard laboratory experiments but also in
274 outdoor perturbation experiments where a VOC is added to a chamber filled with ambient air,
275 control of RO₂ reaction conditions can be challenging.

276 277 *Possible RO₂ fates in chamber experiments of SOA formation*

278
279 A key question is what chemical conditions can be accessed in laboratory chamber experiments
280 measuring SOA yields and to what extent chambers can span the range of RO₂ reaction conditions
281 found in the global atmosphere (as shown in Figure 2). Using a box model with near-explicit
282 chemistry (F0AM (68) with MCM v3.3.1 (69, 70)), we assess the RO₂ fate distribution accessible
283 during photochemical isoprene oxidation in a typical environmental chamber at 298 K, as shown
284 in Figure 5. We span combinations of initial concentrations of H₂O₂, HONO, and NO with 100 ppb
285 initial isoprene, a mixing ratio sufficiently high to allow for SOA yield measurements. For each
286 simulated experiment, we determine whether the initial conditions allow for two generations of
287 oxidation during an 8-hour experiment, and for those experiments that satisfy this criterion, we
288 assess the corresponding distribution of RO₂ fates. The results we present here use experimental
289 parameters (dilution rates, light intensities) from a single chamber (the 7.5 m³ MIT environmental
290 chamber (71)), but the overall conclusions are applicable to most indoor environmental chambers.
291 The challenges presented here would be exacerbated in smaller-volume chambers (including OFRs)
292 because the timescales for dilution and wall loss are shorter, necessitating even higher oxidant
293 concentrations.

294
295 Figure 5 shows results from these box-modeling simulations. The entire range of β between 0 and
296 1 can be achieved by careful selection of precursor concentrations. Achieving atmospheric
297 RO₂/HO₂ ratios is possible at intermediate and high β , but is more difficult at low β , when very

298 high quantities of H₂O₂ are required to achieve sufficiently fast oxidation (72). However, RO₂
299 speciation also differs between chambers and the atmosphere: CH₃O₂ is ≈ 50% of atmospheric RO₂
300 whereas it only makes up ≈ 1-10% of chamber RO₂ (Figure S3). As such, chamber RO₂+RO₂
301 product distributions likely differ from those in the atmosphere. NO/NO₂ ratios are generally lower
302 in chamber conditions than in the atmosphere as a result of high HO₂ concentrations (see Figure
303 S4) and low NO₂ photolysis rates (j_{NO_2}). While increasing the chamber light intensity would result
304 in increased NO/NO₂ ratios, the high chamber HO₂ concentrations preclude reaching atmospheric
305 NO/NO₂ ratios even at atmospheric j_{NO_2} .

306
307 With various initial conditions, bimolecular lifetimes (τ_{bi}) between 2 and 20 seconds can be
308 achieved at low β , but are orders of magnitude lower at high β . At moderate β , experimental
309 conditions can access τ_{bi} in the 10s of seconds, but the lower oxidant levels at these longer τ_{bi} mean
310 it becomes increasingly difficult to achieve two generations of oxidation in an 8-hour experiment.
311 However, in order to capture the atmospheric distribution of bimolecular and unimolecular RO₂
312 products in a chamber experiment, it is not always necessary to have values of τ_{bi} in the chamber
313 directly overlap with those in the atmosphere. Instead, experiments need only to access a regime
314 where the relative importance of bimolecular and unimolecular reactions is largely the same as it is
315 in the atmosphere. Accessing atmospherically relevant RO₂ fates for RO₂ with fast or slow
316 isomerization ($\tau_{\text{uni}} < 1$ s or $\tau_{\text{uni}} > 1,000$ s) is achievable with many combinations of initial conditions
317 in the chamber. Careful selection of initial conditions is required, however, for RO₂ with
318 intermediate isomerization rates (1 s $< \tau_{\text{uni}} < 1,000$ s) to ensure that competition between
319 bimolecular and unimolecular fates in the chamber match that of the atmosphere.

320
321 In the case of isoprene, eight different first-generation RO₂ radicals are formed, but only two (the
322 Z- δ isomers) undergo isomerization reactions at atmospheric conditions. Both Z- δ isomers
323 isomerize rapidly enough (4-OH Z- δ $\tau_{\text{uni}} = 0.3$ s, 1-OH Z- δ $\tau_{\text{uni}} = 3$ s at 297 K (55)) that
324 isomerization almost always outcompetes their bimolecular reactions in the atmosphere. SOA
325 chamber experiments can mimic atmospheric conditions where the fates of these RO₂ are
326 dominated by isomerization across the entire β range for the 4-OH Z- δ isomer and at high- and
327 intermediate- β for the 1-OH Z- δ isomer. Low- β experiments can achieve conditions in which
328 isomerization outcompetes bimolecular reactions of the 1-OH Z- δ RO₂ but cannot access
329 atmospheric conditions where isomerization entirely *dominates* the 1-OH Z- δ RO₂ fate. We note
330 that β and τ_{bi} derived from GEOS-Chem shown in Figure 5 are calculated at the corresponding
331 atmospheric temperatures and pressures, whereas the β and τ_{bi} that describe chamber chemistry are
332 all calculated at 298 K. Over the range of atmospheric isoprene oxidation temperatures, however,
333 variation in τ_{uni} for isoprene-derived RO₂ is constrained to within an order of magnitude (shaded
334 rectangles in Figure 5) and therefore does not significantly alter the atmospheric competition
335 between bimolecular and unimolecular fates for isoprene-derived RO₂.

336
337 Ratios of oxidants also vary between extreme- β experiments and intermediate- β experiments. As
338 shown in Figure S5, extreme- β experiments can access conditions in which nearly all the isoprene
339 is oxidized by OH, without contributions from other oxidants (O₃ and NO₃). At $\beta=0$, the lack of
340 NO prevents production of O₃ or NO₃, while at $\beta=1$, there is sufficient excess of NO to titrate away
341 any O₃ and NO₃ produced. However, at intermediate- β conditions, photochemical cycling will
342 produce O₃ and NO₃, which will persist because the levels of NO are too low to titrate them. In fact,
343 intermediate- β experiments involve OH/O₃ ratios that are similar to those in regions of the
344 atmosphere characterized by intermediate- β conditions. NO₃ concentrations at intermediate- β ,
345 however, will be higher in most indoor chambers compared to daytime tropospheric conditions
346 because of the lower visible light intensities. The presence of multiple oxidants does complicate
347 experimental interpretation and subsequent model parameterization since most current model SOA

parameterizations are based on yields for a single oxidant. However, the SOA produced during intermediate- β experiments is arguably more representative of atmospheric photochemical SOA production than single-oxidant studies.

Changes in chemical conditions during an experiment resulting from the oxidation of both NO_x and VOCs are another consideration during intermediate- β experiments (73). The decreases in NO concentrations over time lead to reductions in β and increases in τ_{bi} . As shown in Figure S6, changes in τ_{bi} over a typical experiment are on the order of seconds to tens of seconds. Changes in β vary and are minimized at the extremes, but intermediate- β experiments can transverse nearly the full range of β values over time, and these time-dependent changes in chemical conditions make it more challenging to locate these experiments in chemical space. Addition of a small amount of NO (or NO precursor) throughout an experiment (28, 64, 74) may reduce the time-dependent changes in chemical conditions.

Chamber experiments spanning a range of RO_2 fates

Using the global model- and box model-informed approach to experimental design described here, we carried out a series of isoprene+OH oxidation chamber experiments. In Figure 6, we focus on three experiments that span the chamber-accessible β - τ_{bi} parameter space. These proof-of-concept experiments are, to our knowledge, the first that explicitly and systematically attempt to span the atmospheric chemical environment during SOA production, despite the aforementioned challenges. Initial conditions for these experiments (shown in Table S1) are used as box modeling inputs to predict RO_2 fate parameters (β and τ_{bi}) for each experiment. To assess whether we achieved the predicted ranges of β and τ_{bi} in these experiments, we examine the first-generation gas-phase oxidative product distribution, shown in Figure 6 and Figure S9. At low- β conditions ($\beta=0.02$), the first-generation product distribution is dominated, as expected, by $\text{C}_5\text{H}_{10}\text{O}_3$; this corresponds to ISOPOOH, the main product formed from RO_2+HO_2 chemistry. At intermediate- and high- β conditions ($\beta=0.61$ and 1.0 , respectively), isoprene hydroxy nitrate ($\text{C}_5\text{H}_9\text{O}_4\text{N}$, the RO_2+NO termination product) becomes an increasingly larger contributor to the product distribution. The relative importance of the isomerization product (HPALD, $\text{C}_5\text{H}_8\text{O}_3$) to the product distribution is highest at intermediate- β conditions in which τ_{bi} is expected to be highest. These first-generation gas-phase product distribution trends confirm that our experiments did indeed span ranges of β and τ_{bi} , demonstrating that chamber experiments can be run under a wider range of atmospheric RO_2 fates than have been accessed previously.

The limitations of our experimental instrumentation preclude measurements of unique RO_2+RO_2 and RO_2+NO_2 products. However, modeling of our experiments indicates that the relative contribution of these pathways to RO_2 fate in our experiments is in line with what is shown in Figure 5: RO_2/HO_2 ratios fall within the atmospheric RO_2/HO_2 distribution whereas NO/NO_2 ratios are far below the atmospheric NO/NO_2 distribution. As a result of the high HO_x concentrations in typical chamber SOA experiments, faster photolysis rates or higher NO concentrations than those under typical atmospheric photochemical conditions are required to achieve atmospheric NO/NO_2 ratios.

Discussion

Pathways forward

The results of the proof-of-concept experiments described above confirm the possibility of designing and executing experiments to mimic many of the key features of the atmospheric chemical environment in laboratory chambers, including chemical environments where multiple

398 RO₂ fates compete (e.g., the intermediate-β, longer-τ_{bi} experiment in Figure 6). As such, this model-
399 informed approach to experimental design represents a useful approach for future chamber
400 experiments of SOA production. The next generation of SOA chamber studies and model
401 treatments of SOA production, however, must find ways to tackle the remaining challenges. Here,
402 we suggest some directions for future research aimed at enabling progress in our understanding of
403 SOA formation chemistry.

404
405 As discussed earlier, one of the key limitations of SOA chamber experiments is the need to produce
406 enough products such that they surpass instrument limits of detection. This is particularly
407 challenging for VOCs with low SOA yields, such as isoprene. The ability to measure SOA mass
408 concentrations with ng/m³-level sensitivity would relax this constraint, allowing experiments to be
409 run with lower initial VOC concentrations. For example, as shown in Figure S7, longer τ_{bi} are
410 accessible during experiments with lower initial isoprene concentrations, including τ_{bi} up to 200 s
411 for high-β experiments initialized with 0.1 ppb isoprene, because of lower HO₂ and RO₂
412 concentrations. Such improvements to analytical instrumentation for measuring OA mass,
413 especially for experiments with added seed particles, could involve increased sensitivity for hard-
414 ionization quantitative aerosol mass measurements (e.g., the aerosol mass spectrometer or aerosol
415 chemical speciation monitor) as well as the development of sensitive (non-mass spectrometric)
416 approaches for measuring OA mass. Better quantification methods for highly sensitive soft
417 ionization techniques (e.g., chemical ionization mass spectrometry and extractive electrospray
418 ionization mass spectrometry) could also help relax this constraint.

419
420 Quantifying SOA yields through new particle formation (NPF) could alleviate some of the LOD
421 constraint by improving signal-to-noise ratios. However, ultrafine particles produced during
422 nucleation events are lost to chamber walls more quickly than the accumulation mode particles
423 typically present during seeded experiments. Therefore, NPF experiments that accommodate multi-
424 generation chemistry require large chambers and fast oxidation, resulting in increased HO_x
425 concentrations and thus altering the radical balance.

426
427 In addition, inclusion of the role of all RO₂ reactions involved in SOA formation may require more
428 complex SOA parameterizations in models. The untested linear mixing assumption between
429 RO₂+NO and RO₂+HO₂ yields should at least be tested, and SOA parameterizations may need to
430 increase in complexity to account for more than two RO₂ fates. For example, existing RO₂-fate-
431 based SOA parameterizations that include SOA yields for RO₂+HO₂ and RO₂+NO (37) may need
432 to include yields for other RO₂ reaction pathways. Future experimental work is required, however,
433 to evaluate how much a given RO₂ reaction pathway actually affects atmospheric SOA production
434 and therefore how many dimensions are required to accurately model SOA formation chemistry. If
435 SOA yields do not vary significantly over the atmospheric ranges of an RO₂ fate, that reaction
436 pathway does not need to be carefully studied or included in models. For example, variation within
437 the atmospheric distribution of τ_{bi} are likely to matter only for systems with isomerization reactions
438 that have lifetimes within an order of magnitude of atmospheric τ_{bi} (i.e., 1 s < τ_{uni} < 1,000 s). Recent
439 work has demonstrated that SOA yields from some monoterpene reactions are sensitive to RO₂/HO₂
440 only when this ratio is changed dramatically (47, 75), suggesting that the atmospheric variation of
441 RO₂/HO₂ between 0 and 2 has an insignificant effect on biogenic SOA yields. Other laboratory
442 studies, however, have suggested that this narrow atmospheric range may affect SOA yields from
443 anthropogenic VOCs (46). Additional experiments have shown that isoprene SOA yields more than
444 halve between NO/NO₂ ≈ 0.1 and 0.3 (44), suggesting that changes in NO/NO₂ across the range
445 found in the atmosphere (≈ 0.2-0.7) can affect isoprene SOA yields. Similar tests are required across
446 these atmospheric ranges for an assortment of VOCs.

448 An alternative to the parameterized approach for modeling SOA production is a semi-mechanistic
449 approach, which describes SOA formation in terms of key condensable species generated from gas-
450 phase chemical mechanisms. This approach combines laboratory measurements of gas-phase
451 products with estimates of their volatilities to predict particle-phase products and aerosol yields
452 and was recently implemented in the Community Regional Atmospheric Chemistry Multiphase
453 Mechanism (CRACMM) (76). Such an approach has the advantage of inherently including the
454 effects of changes in atmospheric variables (e.g., RO₂ chemistry, temperature) on SOA production,
455 in contrast to a parameterized approach where such effects require explicit consideration. However,
456 a semi-mechanistic approach generally requires uncertain estimates of oxidation product volatilities
457 as well as thoughtful consideration of what species are included given resource constraints, and
458 predicted SOA formation has generally not been validated by yields measured in laboratory studies.
459 More model species will also increase computational demand, and so may not be appropriate for
460 large-scale models, including global chemistry-climate and/or Earth system models.

461
462 Regardless of the model approach used to describe SOA formation, a crucial component is
463 continual interfacing between laboratory measurements, models, and field observations. Updated
464 laboratory-based model predictions of SOA should be compared to ambient observations, and
465 further experiments should then be targeted to chemical environments in which the model-
466 measurement difference is greatest. Validation of model SOA descriptions against SOA yields
467 measured in chamber experiments that match the atmospheric chemical environment will also aid
468 model parameterization and further development.

469 *Conclusions*

470
471
472 The mechanistic advances in our understanding of RO₂ chemistry over the last two decades have
473 enabled us to replace the high- and low-NO_x binary used in the past to describe RO₂ fates with a
474 more nuanced description of the complex, multi-dimensional space that controls the product
475 distribution of atmospheric VOC oxidation. Leveraging 3D- and box-modeling tools with our
476 understanding of the complexity of RO₂ chemistry along with careful control of chamber inputs
477 will help ensure that experimental resources are dedicated to experiments that span chemical
478 parameter spaces that best match those of the atmosphere.

479
480 Because laboratory studies are the foundation for model parameterizations of SOA, their limitations
481 inherently become the limitations of models. Modeling SOA based on experiments run at extreme-
482 β conditions has major shortcomings: extreme-β conditions do not describe the chemical
483 environment in most of the atmosphere, the assumption of linearity between extremes has never
484 been tested, and extreme-β chamber conditions generally do not mimic the role of other RO₂
485 reactions (RO₂ isomerization, RO₂+RO₂, and RO₂+NO₂) that may occur in the atmosphere.
486 However, improvements are possible, via the use of modeling to explicitly design the chemical
487 environment for SOA chamber experiments. The fact that models are fundamentally limited by
488 laboratory constraints also means that model results should be interpreted with the limitations of
489 chamber experiments in mind. The work presented here indicates that some atmospheric RO₂ fate
490 distributions can be better represented in chambers than others. For example, achieving τ_{bi} that
491 allows for the same competition between bimolecular and unimolecular RO₂ reactions as what
492 occurs in the atmosphere is most challenging for low-β environments, but can be reasonably well-
493 captured under more polluted high-β conditions. As such, interpretation of model results, and their
494 ability to reproduce field observations, should be done with awareness of the relevant RO₂ chemical
495 environment.

497 Experimental chamber work over the last several decades has advanced our understanding of the
498 complexity of the multi-generation oxidative chemistry involved in VOC oxidation and SOA
499 production. To achieve the ultimate goal of fully understanding the processes that control the
500 spatiotemporal distribution of pollutants in the atmosphere, the next generation of chamber studies
501 must be more tightly connected to true atmospheric reaction conditions of RO₂ radicals. Such
502 studies are the essential foundation for interpreting past and future field observations, and for
503 building the accurate model parameterizations that are required to fully understand and predict
504 atmospheric SOA, and ultimately its impacts on air quality and climate.

506 **Materials and Methods**

507 *Modeling*

508
509 We use the GEOS-Chem chemical transport model (version 13.4.0) at a horizontal grid resolution
510 of 2°x2.5° and with 47 vertical layers. Results shown are from simulations of January and July 2016
511 after a 1-year spin-up with output saved every hour. The model is driven by assimilated
512 meteorology from the Modern-Era Retrospective analysis for Research and Applications, Version
513 2 (MERRA-2), from the NASA Global Modeling and Assimilation Office (GMAO). Our
514 simulations use the full gas-phase chemistry available in GEOS-Chem, including a coupled HO_x-
515 NO_x-VOC-O₃-halogen chemical mechanism (77–80). Biogenic emissions are calculated online
516 using the MEGANv2.1 framework (81) and anthropogenic emissions are from the global CEDS
517 inventory (82).

518
519 Box model simulations were run using the Framework for 0D Atmospheric Modeling (F0AM) (68)
520 with near-explicit chemistry described by the Master Chemical Mechanism (MCM v3.3.1) (69, 70).
521 Our simulations use a measured light spectrum scaled by a measured j_{NO_2} rate of 0.12 min⁻¹. In
522 order to span possible chamber initial conditions, we run a series of 8-hour simulations where we
523 logarithmically permute initial isoprene (4 values between 0.1 and 100 ppb, inclusive, shown in
524 Figure S7), HONO (18 values between 0 and 1,000 ppb, inclusive), H₂O₂ (12 values between 10
525 and 10,000 ppb, inclusive), and NO (18 values between 0 and 1,000 ppb, inclusive) concentrations
526 that vary over orders of magnitude (Figures 5 and S8). We run a separate series of simulations
527 where we permute initial CH₃ONO and NO concentrations (Figure S8). We consider a simulation
528 in which the final methyl vinyl ketone (MVK) concentration is less than half of the maximum MVK
529 concentration as an experiment that achieved second generation oxidation chemistry. For each
530 simulation, we calculate β , τ_{bi} , RO₂/HO₂ and NO/NO₂. For experiments that achieve second-
531 generation chemistry, the experiment end is defined as the time at which the MVK concentration
532 decreases to half of its maximum. Other experiments are assumed to end at 8 hours. β is calculated
533 from the cumulative integrated rates of RO₂ loss over the course of the experiment; τ_{bi} , RO₂/HO₂,
534 and NO/NO₂ are calculated from the mean value during the simulated experiment.

536 *Chamber experiments*

537
538 Experiments were conducted in a 7.5 m³ temperature-controlled environmental chamber (71) held
539 at 20°C and RH < 10%. The chamber is constructed of PFA Teflon and is surrounded by 48 UV
540 lamps whose emission is centered at 340 nm and whose intensity drives NO₂ photolysis at a rate of
541 0.12 min⁻¹. The chamber is operated in ‘semi-batch’ mode in which clean air is continuously added
542 to make up for instrument sample flow to maintain a constant volume. Between experiments, the
543 chamber is flushed by zero air for at least 12 h to ensure a clean background. Prior to the start of
544 each experiment, ammonium sulfate seed particles are injected to serve as condensation nuclei,
545 followed by injection of acetonitrile (dilution tracer), isoprene, and the oxidant precursors (HONO
546 and/or H₂O₂). Seed particles are atomized from a 2 g/L solution of ammonium sulfate in water.

547 Isoprene and H₂O₂ are added through a silicone septum into a 10 lpm flow. HONO is generated by
548 adding sulfuric acid via a syringe to a constantly stirred solution of sodium nitrite; the HONO
549 produced, as well as any NO and NO₂ produced as co-products, are carried into the chamber via a
550 1 lpm air stream. Experiments are initiated by turning on the chamber lights.

551

552 The gas-phase composition of the reaction mixture was monitored with a proton-transfer-reaction
553 time-of-flight mass spectrometer (Vocus PTR-MS, Aerodyne Research Inc.) (83) and an
554 ammonium time-of-flight chemical ionization mass spectrometer (NH₄⁺ CIMS, Ionicon Analytic)
555 (84). Though not discussed here, particle-phase measurements were also made with an aerosol mass
556 spectrometer (AMS, Aerodyne Research Inc.) and a scanning mobility particle sizer (TSI).
557 Additional measurements include those from a NO-NO₂-NO_x analyzer (Thermo Scientific), a
558 Cavity Attenuated Phase Shift Spectroscopy NO₂ monitor (CAPS NO₂) (Aerodyne Research Inc.),
559 and an ozone monitor (2B Technologies).

560

561 Parameters that describe RO₂ fate (β , τ_{bi} , RO₂/HO₂, and NO/NO₂) for these experiments are
562 calculated using MCM v3.3.1 box model simulations in F0AM as described above. Simulations are
563 initialized with initial measured NO (from the NO channel of the Thermo Scientific NO-NO₂-NO_x
564 analyzer), NO₂ (from the CAPS NO₂ monitor), HONO (estimated by subtracting the CAPS NO₂
565 signal from the NO_x analyzer NO₂ signal), and isoprene concentrations (from the Vocus PTR-MS).
566 Initial H₂O₂ concentrations were estimated based on injection amount.

567

568

569 References

- 570 1. R. Burnett, H. Chen, M. Szyszkowicz, N. Fann, B. Hubbell, C. A. Pope, J. S. Apte, M.
571 Brauer, A. Cohen, S. Weichenthal, J. Coggins, Q. Di, B. Brunekreef, J. Frostad, S. S. Lim,
572 H. Kan, K. D. Walker, G. D. Thurston, R. B. Hayes, C. C. Lim, M. C. Turner, M. Jerrett, D.
573 Krewski, S. M. Gapstur, W. R. Diver, B. Ostro, D. Goldberg, D. L. Crouse, R. V. Martin, P.
574 Peters, L. Pinault, M. Tjepkema, A. van Donkelaar, P. J. Villeneuve, A. B. Miller, P. Yin,
575 M. Zhou, L. Wang, N. A. H. Janssen, M. Marra, R. W. Atkinson, H. Tsang, T. Quoc Thach,
576 J. B. Cannon, R. T. Allen, J. E. Hart, F. Laden, G. Cesaroni, F. Forastiere, G. Weinmayr, A.
577 Jaensch, G. Nagel, H. Concin, J. V. Spadaro, Global estimates of mortality associated with
578 long-term exposure to outdoor fine particulate matter. *Proceedings of the National Academy
579 of Sciences* **115**, 9592–9597 (2018).
- 580 2. C. J. L. Murray, A. Y. Aravkin, P. Zheng, C. Abbafati, K. M. Abbas, M. Abbasi-Kangevari,
581 F. Abd-Allah, A. Abdelalim, M. Abdollahi, I. Abdollahpour, K. H. Abegaz, H. Abolhassani,
582 V. Aboyans, L. G. Abreu, M. R. M. Abrigo, A. Abualhasan, L. J. Abu-Raddad, A. I.
583 Abushouk, M. Adabi, V. Adekanmbi, A. M. Adeoye, O. O. Adetokunboh, D. Adham, S. M.
584 Advani, G. Agarwal, S. M. K. Aghamir, A. Agrawal, T. Ahmad, K. Ahmadi, M. Ahmadi, H.
585 Ahmadieh, M. B. Ahmed, T. Y. Akalu, R. O. Akinyemi, T. Akinyemiju, B. Akombi, C. J.
586 Akunna, F. Alahdab, Z. Al-Aly, K. Alam, S. Alam, T. Alam, F. M. Alanezi, T. M. Alanzi, B.
587 wassihun Alemu, K. F. Alhabib, M. Ali, S. Ali, G. Alicandro, C. Alinia, V. Alipour, H.
588 Alizade, S. M. Aljunid, F. Alla, P. Allebeck, A. Almasi-Hashiani, H. M. Al-Mekhlafi, J.
589 Alonso, K. A. Altirkawi, M. Amini-Rarani, F. Amiri, D. A. Amugsi, R. Ancuceanu, D.
590 Anderlini, J. A. Anderson, C. L. Andrei, T. Andrei, C. Angus, M. Anjomshoa, F. Ansari, A.
591 Ansari-Moghaddam, I. C. Antonazzo, C. A. T. Antonio, C. M. Antony, E. Antriyandarti, D.
592 Anvari, R. Anwer, S. C. Y. Appiah, J. Arabloo, M. Arab-Zozani, F. Ariani, B. Armoon, J.
593 Ärnlov, A. Arzani, M. Asadi-Aliabadi, A. A. Asadi-Pooya, C. Ashbaugh, M. Assmus, Z.
594 Atafar, D. D. Atnafu, M. M. W. Atout, F. Ausloos, M. Ausloos, B. P. A. Quintanilla, G.
595 Ayano, M. A. Ayanore, S. Azari, G. Azarian, Z. N. Azene, A. Badawi, A. D. Badiye, M. A.
596 Bahrami, M. H. Bakhshaei, A. Bakhtiari, S. M. Bakkannavar, A. Baldasseroni, K. Ball, S. H.

597 Ballew, D. Balzi, M. Banach, S. K. Banerjee, A. B. Bante, A. G. Baraki, S. L. Barker-Collo,
598 T. W. Bärnighausen, L. H. Barrero, C. M. Barthelemy, L. Barua, S. Basu, B. T. Baune, M.
599 Bayati, J. S. Becker, N. Bedi, E. Beghi, Y. Béjot, M. L. Bell, F. B. Bennitt, I. M. Bensor,
600 K. Berhe, A. E. Berman, A. S. Bhagavathula, R. Bhageerathy, N. Bhala, D. Bhandari, K.
601 Bhattacharyya, Z. A. Bhutta, A. Bijani, B. Bikbov, M. S. B. Sayeed, A. Biondi, B. M.
602 Birihane, C. Bisignano, R. K. Biswas, H. Bitew, S. Bohlouli, M. Bohluli, A. S. Boon-
603 Dooley, G. Borges, A. M. Borzi, S. Borzouei, C. Bosetti, S. Boufous, D. Braithwaite, N. J.
604 K. Breitborde, S. Breitner, H. Brenner, P. S. Briant, A. N. Briko, N. I. Briko, G. B. Britton,
605 D. Bryazka, B. R. Bumgarner, K. Burkart, R. T. Burnett, S. B. Nagaraja, Z. A. Butt, F. L. C.
606 dos Santos, L. E. Cahill, L. L. A. Cámera, I. R. Campos-Nonato, R. Cárdenas, G. Carreras, J.
607 J. Carrero, F. Carvalho, J. M. Castaldelli-Maia, C. A. Castañeda-Orjuela, G. Castelpietra, F.
608 Castro, K. Causey, C. R. Cederroth, K. M. Cercy, E. Cerin, J. S. Chandan, K.-L. Chang, F. J.
609 Charlson, V. K. Chattu, S. Chaturvedi, N. Cherbuin, O. Chimed-Ochir, D. Y. Cho, J.-Y. J.
610 Choi, H. Christensen, D.-T. Chu, M. T. Chung, S.-C. Chung, F. M. Cicutini, L. G. Ciobanu,
611 M. Cirillo, T. K. D. Classen, A. J. Cohen, K. Compton, O. R. Cooper, V. M. Costa, E.
612 Cousin, R. G. Cowden, D. H. Cross, J. A. Cruz, S. M. A. Dahlawi, A. A. M. Damasceno, G.
613 Damiani, L. Dandona, R. Dandona, W. J. Dangel, A.-K. Danielsson, P. I. Dargan, A. M.
614 Darwesh, A. Daryani, J. K. Das, R. D. Gupta, J. das Neves, C. A. Dávila-Cervantes, D. V.
615 Davitoiu, D. D. Leo, L. Degenhardt, M. DeLang, R. P. Dellavalle, F. M. Demeke, G. T.
616 Demoz, D. G. Demsie, E. Denova-Gutiérrez, N. Derveniz, G. P. Dhungana, M.
617 Dianatinasab, D. D. da Silva, D. Diaz, Z. S. D. Forooshani, S. Djalalinia, H. T. Do, K.
618 Dokova, F. Dorostkar, L. Doshmangir, T. R. Driscoll, B. B. Duncan, A. R. Duraes, A. W.
619 Eagan, D. Edvardsson, N. E. Nahas, I. E. Sayed, M. E. Tantawi, I. Elbarazi, I. Y. Elgendy, S.
620 I. El-Jaafary, I. R. Elyazar, S. Emmons-Bell, H. E. Erskine, S. Eskandarieh, S. Esmacilnejad,
621 A. Esteghamati, K. Estep, A. Etemadi, A. E. Etilso, J. Fanzo, M. Farahmand, M. Fareed, R.
622 Faridnia, A. Farioli, A. Faro, M. Faruque, F. Farzadfar, N. Fattahi, M. Fazlzadeh, V. L.
623 Feigin, R. Feldman, S.-M. Fereshtehnejad, E. Fernandes, G. Ferrara, A. J. Ferrari, M. L.
624 Ferreira, I. Filip, F. Fischer, J. L. Fisher, L. S. Flor, N. A. Foigt, M. O. Folayan, A. A.
625 Fomenkov, L. M. Force, M. Foroutan, R. C. Franklin, M. Freitas, W. Fu, T. Fukumoto, J. M.
626 Furtado, M. M. Gad, E. Gakidou, S. Gallus, A. L. Garcia-Basteiro, W. M. Gardner, B. S.
627 Geberemariam, A. A. A. Gebreslassie, A. Geremew, A. G. Hayoon, P. W. Gething, M.
628 Ghadimi, K. Ghadiri, F. Ghaffarifar, M. Ghafourifard, F. Ghamari, A. Ghashghaee, H.
629 Ghiasvand, N. Ghith, A. Gholamian, R. Ghosh, P. S. Gill, T. G. G. Ginindza, G. Giussani, E.
630 V. Gnedovskaya, S. Goharinezhad, S. V. Gopalani, G. Gorini, H. Goudarzi, A. C. Goulart,
631 F. Greaves, M. Grivna, G. Grosso, M. I. M. Gubari, H. C. Gugnani, R. A. Guimarães, R. A.
632 Guled, G. Guo, Y. Guo, R. Gupta, T. Gupta, B. Haddock, N. Hafezi-Nejad, A. Hafiz, A.
633 Haj-Mirzaian, A. Haj-Mirzaian, B. J. Hall, I. Halvaei, R. R. Hamadeh, S. Hamidi, M. S.
634 Hammer, G. J. Hankey, H. Haririan, J. M. Haro, A. I. Hasaballah, M. M. Hasan, E.
635 Hasanpoor, A. Hashi, S. Hassanipour, H. Hassankhani, R. J. Havmoeller, S. I. Hay, K.
636 Hayat, G. Heidari, R. Heidari-Soureshjani, H. J. Henrikson, M. E. Herbert, C. Herteliu, F.
637 Heydarpour, T. R. Hird, H. W. Hoek, R. Holla, P. Hoogar, H. D. Hosgood, N. Hossain, M.
638 Hosseini, M. Hosseinzadeh, M. Hostiuc, S. Hostiuc, M. Househ, M. Hsairi, V. C. Hsieh, G.
639 Hu, K. Hu, T. M. Huda, A. Humayun, C. K. Huynh, B.-F. Hwang, V. C. Iannucci, S. E.
640 Ibitoye, N. Ikeda, K. S. Ikuta, O. S. Ilesanmi, I. M. Ilic, M. D. Ilic, L. R. Inbaraj, H. Ippolito,
641 U. Iqbal, S. S. N. Irvani, C. M. S. Irvine, M. M. Islam, S. M. S. Islam, H. Iso, R. Q. Ivers, C.
642 C. D. Iwu, C. J. Iwu, I. O. Iyamu, J. Jaafari, K. H. Jacobsen, H. Jafari, M. Jafarinia, M. A.
643 Jahani, M. Jakovljevic, F. Jalilian, S. L. James, H. Janjani, T. Javaheri, J. Javidnia, P.
644 Jeemon, E. Jenabi, R. P. Jha, V. Jha, J. S. Ji, L. Johansson, O. John, Y. O. John-Akinola, C.
645 O. Johnson, J. B. Jonas, F. Joukar, J. J. Jozwiak, M. Jürisson, A. Kabir, Z. Kabir, H. Kalani,
646 R. Kalani, L. R. Kalankesh, R. Kalhor, T. Kanchan, N. Kapoor, B. K. Matin, A. Karch, M.

647 A. Karim, G. M. Kassa, S. V. Katikireddi, G. A. Kayode, A. K. Karyani, P. N. Keiyoro, C.
648 Keller, L. Kemmer, P. J. Kendrick, N. Khalid, M. Khammarnia, E. A. Khan, M. Khan, K.
649 Khatab, M. M. Khater, M. N. Khatib, M. Khayamzadeh, S. Khazaei, C. Kieling, Y. J. Kim,
650 R. W. Kimokoti, A. Kisa, S. Kisa, M. Kivimäki, L. D. Knibbs, A. K. S. Knudsen, J. M.
651 Kocarnik, S. Kochhar, J. A. Kopec, V. A. Korshunov, P. A. Koul, A. Koyanagi, M. U. G.
652 Kraemer, K. Krishan, K. J. Krohn, H. Kromhout, B. K. Defo, G. A. Kumar, V. Kumar, O. P.
653 Kurmi, D. Kusuma, C. L. Vecchia, B. Lacey, D. K. Lal, R. Laloo, T. Lallukka, F. H. Lami,
654 I. Landires, J. J. Lang, S. M. Langan, A. O. Larsson, S. Lasrado, P. Lauriola, J. V. Lazarus,
655 P. H. Lee, S. W. H. Lee, K. E. LeGrand, J. Leigh, M. Leonardi, H. Lescinsky, J. Leung, M.
656 Levi, S. Li, L.-L. Lim, S. Linn, S. Liu, S. Liu, Y. Liu, J. Lo, A. D. Lopez, J. C. F. Lopez, P.
657 D. Lopukhov, S. Lorkowski, P. A. Lotufo, A. Lu, A. Lugo, E. R. Maddison, P. W. Mahasha,
658 M. M. Mahdavi, M. Mahmoudi, A. Majeed, A. Maleki, S. Maleki, R. Malekzadeh, D. C.
659 Malta, A. A. Mamun, A. L. Manda, H. Manguerra, F. Mansour-Ghanaei, B. Mansouri, M. A.
660 Mansournia, A. M. M. Herrera, J. C. Maravilla, A. Marks, R. V. Martin, S. Martini, F. R.
661 Martins-Melo, A. Masaka, S. Z. Masoumi, M. R. Mathur, K. Matsushita, P. K. Maulik, C.
662 McAlinden, J. J. McGrath, M. McKee, M. M. Mehndiratta, F. Mehri, K. M. Mehta, Z. A.
663 Memish, W. Mendoza, R. G. Menezes, E. W. Mengesha, A. Mereke, S. T. Mereta, A.
664 Meretoja, T. J. Meretoja, T. Mestrovic, B. Miazgowski, T. Miazgowski, I. M. Michalek, T.
665 R. Miller, E. J. Mills, G. K. Mini, M. Miri, A. Mirica, E. M. Mirrakhimov, H. Mirzaei, M.
666 Mirzaei, R. Mirzaei, M. Mirzaei-Alavijeh, A. T. Misganaw, P. Mithra, B. Moazen, D. K.
667 Mohammad, Y. Mohammad, N. M. G. Mezerji, A. Mohammadian-Hafshejani, N.
668 Mohammadifard, R. Mohammadpourhodki, A. S. Mohammed, H. Mohammed, J. A.
669 Mohammed, S. Mohammed, A. H. Mokdad, M. Molokhia, L. Monasta, M. D. Mooney, G.
670 Moradi, M. Moradi, M. Moradi-Lakeh, R. Moradzadeh, P. Moraga, L. Morawska, J.
671 Morgado-da-Costa, S. D. Morrison, A. Mosapour, J. F. Mosser, S. Mouodi, S. M. Mousavi,
672 A. M. Khaneghah, U. O. Mueller, S. Mukhopadhyay, E. C. Mullany, K. I. Musa, S.
673 Muthupandian, A. F. Nabhan, M. Naderi, A. J. Nagarajan, G. Nagel, M. Naghavi, B.
674 Naghshtabrizi, M. D. Naimzada, F. Najafi, V. Nangia, J. R. Nansseu, M. Naserbakht, V. C.
675 Nayak, I. Negoï, J. W. Ngunjiri, C. T. Nguyen, H. L. T. Nguyen, M. Nguyen, Y. T. Nigatu,
676 R. Nikbakhsh, M. R. Nixon, C. A. Nnaji, S. Nomura, B. Norrving, J. J. Noubiap, C. Nowak,
677 V. Nunez-Samudio, A. Oțoiu, B. Oancea, C. M. Odell, F. A. Ogbo, I.-H. Oh, E. W. Okunga,
678 M. Oladnabi, A. T. Olagunju, B. O. Olusanya, J. O. Olusanya, M. O. Omer, K. L. Ong, O. E.
679 Onwujekwe, H. M. Orpana, A. Ortiz, O. Osarenotor, F. B. Osei, S. M. Ostroff, N.
680 Otstavnov, S. S. Otstavnov, S. Øverland, M. O. Owolabi, M. P. A, J. R. Padubidri, R.
681 Palladino, S. Panda-Jonas, A. Pandey, C. D. H. Parry, M. Pasovic, D. K. Pasupula, S. K.
682 Patel, M. Pathak, S. B. Patten, G. C. Patton, H. P. Toroudi, A. E. Peden, A. Pennini, V. C. F.
683 Pepito, E. K. Peprah, D. M. Pereira, K. Pesudovs, H. Q. Pham, M. R. Phillips, C. Piccinelli,
684 T. M. Pilz, M. A. Piradov, M. Pirsaeheb, D. Plass, S. Polinder, K. R. Polkinghorne, C. D.
685 Pond, M. J. Postma, H. Pourjafar, F. Pourmalek, A. Poznańska, S. I. Prada, V. Prakash, D.
686 R. A. Pribadi, E. Pupillo, Z. Q. Syed, M. Rabiee, N. Rabiee, A. Radfar, A. Rafiee, A. Raggi,
687 M. A. Rahman, A. Rajabpour-Sanati, F. Rajati, I. Rakovac, P. Ram, K. Ramezanzadeh, C. L.
688 Ranabhat, P. C. Rao, S. J. Rao, V. Rashedi, P. Rathi, D. L. Rawaf, S. Rawaf, L. Rawal, R.
689 Rawassizadeh, R. Rawat, C. Razo, S. B. Redford, R. C. Reiner, M. B. Reitsma, G. Remuzzi,
690 V. Renjith, A. M. N. Renzaho, S. Resnikoff, N. Rezaei, N. Rezaei, A. Rezapour, P.-A.
691 Rhinehart, S. M. Riahi, D. C. Ribeiro, D. Ribeiro, J. Rickard, J. A. Rivera, N. L. S. Roberts,
692 S. Rodríguez-Ramírez, L. Roeber, L. Ronfani, R. Room, G. Roshandel, G. A. Roth, D.
693 Rothenbacher, E. Rubagotti, G. M. Rweggerera, S. Sabour, P. S. Sachdev, B. Saddik, E.
694 Sadeghi, M. Sadeghi, R. Saeedi, S. S. Moghaddam, Y. Safari, S. Safi, S. Safiri, R. Sagar, A.
695 Sahebkar, S. M. Sajadi, N. Salam, P. Salamati, H. Salem, M. R. R. Salem, H. Salimzadeh, O.
696 M. Salman, J. A. Salomon, Z. Samad, H. S. Kafil, E. Z. Sambala, A. M. Samy, J. Sanabria,

- 697 T. G. Sánchez-Pimienta, D. F. Santomauro, I. S. Santos, J. V. Santos, M. M. Santric-
698 Milicevic, S. Y. I. Saraswathy, R. Sarmiento-Suárez, N. Sarrafzadegan, B. Sartorius, A.
699 Sarveazad, B. Sathian, T. Sathish, D. Sattin, S. Saxena, L. E. Schaeffer, S. Schiavolin, M. P.
700 Schlaich, M. I. Schmidt, A. E. Schutte, D. C. Schwebel, F. Schwendicke, A. M. Senbeta, S.
701 Senthilkumaran, S. G. Sepanlou, B. Serdar, M. L. Serre, J. Shadid, O. Shafaat, S. Shahabi,
702 A. A. Shaheen, M. A. Shaikh, A. S. Shalash, M. Shams-Beyranvand, M. Shamsizadeh, K.
703 Sharafi, A. Sheikh, A. Sheikhtaheri, K. Shibuya, K. D. Shield, M. Shigematsu, J. I. Shin, M.-
704 J. Shin, R. Shiri, R. Shirkoohi, K. Shuval, S. Siabani, R. Sierpinski, I. D. Sigfusdottir, R.
705 Sigurvinsdottir, J. P. Silva, K. E. Simpson, J. A. Singh, P. Singh, E. Skiadaresi, S. T. Skou,
706 V. Y. Skryabin, E. U. R. Smith, A. Soheili, S. Soltani, M. Soofi, R. J. D. Sorensen, J. B.
707 Soriano, M. B. Sorrie, S. Soshnikov, I. N. Soyiri, C. N. Spencer, A. Spotin, C. T.
708 Sreeramareddy, V. Srinivasan, J. D. Stanaway, C. Stein, D. J. Stein, C. Steiner, L. Stockfelt,
709 M. A. Stokes, K. Straif, J. L. Stubbs, M. B. Sufiyan, H. A. R. Suleria, R. S. Abdulkader, G.
710 Sulo, I. Sultan, Ł. Szumowski, R. Tabarés-Seisdedos, K. M. Tabb, T. Tabuchi, A.
711 Taherkhani, M. Tajdini, K. Takahashi, J. S. Takala, A. T. Tamiru, N. Taveira, A. Tehrani-
712 Banihashemi, M.-H. Temsah, G. A. Tesema, Z. T. Tessema, G. D. Thurston, M. V. Titova,
713 H. R. Tohidinik, M. Tonelli, R. Topor-Madry, F. Topouzis, A. E. Torre, M. Touvier, M. R.
714 R. Tovani-Palone, B. X. Tran, R. Travillian, A. Tsatsakis, L. T. Car, S. Tyrovolas, R. Uddin,
715 C. D. Umeokonkwo, B. Unnikrishnan, E. Upadhyay, M. Vacante, P. R. Valdez, A. van
716 Donkelaar, T. J. Vasankari, Y. Vasseghian, Y. Veisani, N. Venketasubramanian, F. S.
717 Violante, V. Vlassov, S. E. Vollset, T. Vos, R. Vukovic, Y. Waheed, M. T. Wallin, Y.
718 Wang, Y.-P. Wang, A. Watson, J. Wei, M. Y. W. Wei, R. G. Weintraub, J. Weiss, A.
719 Werdecker, J. J. West, R. Westerman, J. L. Whisnant, H. A. Whiteford, K. E. Wiens, C. D.
720 A. Wolfe, S. S. Wozniak, A.-M. Wu, J. Wu, S. W. Hanson, G. Xu, R. Xu, S. Yadgir, S. H.
721 Y. Jabbari, K. Yamagishi, M. Yaminfirooz, Y. Yano, S. Yaya, V. Yazdi-Feyzabadi, T. Y.
722 Yeheyis, C. S. Yilgwan, M. T. Yilma, P. Yip, N. Yonemoto, M. Z. Younis, T. P. Younker,
723 B. Yousefi, Z. Yousefi, T. Yousefinezhadi, A. Y. Yousuf, C. Yu, H. Yusefzadeh, T. Z.
724 Moghadam, M. Zamani, M. Zamanian, H. Zandian, M. S. Zastrozhin, Y. Zhang, Z.-J.
725 Zhang, J. T. Zhao, X.-J. G. Zhao, Y. Zhao, M. Zhou, A. Ziapour, S. R. M. Zimsen, M.
726 Brauer, A. Afshin, S. S. Lim, Global burden of 87 risk factors in 204 countries and
727 territories, 1990–2019: a systematic analysis for the Global Burden of Disease Study 2019.
728 *The Lancet* **396**, 1223–1249 (2020).
- 729 3. IPCC, “Climate Change 2021: The Physical Science Basis. Contribution of Working Group
730 I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change”
731 (Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA,
732 2021); 10.1017/9781009157896.
- 733 4. Q. Zhang, J. L. Jimenez, M. R. Canagaratna, J. D. Allan, H. Coe, I. Ulbrich, M. R. Alfarra,
734 A. Takami, A. M. Middlebrook, Y. L. Sun, K. Dzepina, E. Dunlea, K. Docherty, P. F.
735 Decarlo, D. Salcedo, T. Onasch, J. T. Jayne, T. Miyoshi, A. Shimono, S. Hatakeyama, N.
736 Takegawa, Y. Kondo, J. Schneider, F. Drewnick, L. Cottrell, R. J. Griffin, J. Rautiainen, J.
737 Y. Sun, Y. M. Zhang, Ubiquity and dominance of oxygenated species in organic aerosols in
738 anthropogenically-influenced Northern Hemisphere midlatitudes. *Geophysical Research*
739 *Letters* **34**, 1–6 (2007).
- 740 5. B. J. Turpin, J. J. Huntzicker, Identification of secondary organic aerosol episodes and
741 quantitation of primary and secondary organic aerosol concentrations during SCAQS.
742 *Atmospheric Environment* **29**, 3527–3544 (1995).

- 743 6. M. Shrivastava, C. D. Cappa, J. Fan, Allen. H. Goldstein, A. B. Guenther, J. L. Jimenez, C.
744 Kuang, A. Laskin, S. T. Martin, N. L. Ng, T. Petaja, J. R. Pierce, P. J. Rasch, P. Roldin, J. H.
745 Seinfeld, J. Shilling, J. N. Smith, J. A. Thornton, R. Volkamer, J. Wang, D. R. Worsnop, R.
746 A. Zaveri, A. Zelenyuk, Q. Zhang, Recent advances in understanding secondary organic
747 aerosol: Implications for global climate forcing. *Reviews of Geophysics* **55**, 509–559 (2017).
- 748 7. J. Tyndall, XIX. On the action of rays of high refrangibility upon gaseous matter. *Phil.*
749 *Trans. R. Soc.* **160**, 333–365 (1870).
- 750 8. A. J. Haagen-Smit, Chemistry and physiology of Los Angeles smog. *Ind. Eng. Chem.;*
751 *(United States)* **44:6** (1952).
- 752 9. F. W. Went, Blue Hazes in the Atmosphere. *Nature* **187**, 641–643 (1960).
- 753 10. R. M. Kamens, H. E. Jeffries, M. W. Gery, R. W. Wiener, K. G. Sexton, G. B. Howe, The
754 impact of α -pinene on urban smog formation: an outdoor smog chamber study. *Atmospheric*
755 *Environment (1967)* **15**, 969–981 (1981).
- 756 11. S. Hatakeyama, K. Izumi, T. Fukuyama, H. Akimoto, Reactions of ozone with α -pinene and
757 β -pinene in air: Yields of gaseous and particulate products. *Journal of Geophysical*
758 *Research: Atmospheres* **94**, 13013–13024 (1989).
- 759 12. K. Izumi, T. Fukuyama, Photochemical aerosol formation from aromatic hydrocarbons in
760 the presence of NO_x. *Atmospheric Environment. Part A. General Topics* **24**, 1433–1441
761 (1990).
- 762 13. S. E. Paulson, S. N. Pandis, U. Baltensperger, J. H. Seinfeld, R. C. Flagan, E. J. Palen, D. T.
763 Allen, C. Schaffner, W. Giger, A. Portmann, Characterization of photochemical aerosols
764 from biogenic hydrocarbons. *Journal of Aerosol Science* **21**, S245–S248 (1990).
- 765 14. S. N. Pandis, S. E. Paulson, J. H. Seinfeld, R. C. Flagan, Aerosol formation in the
766 photooxidation of isoprene and β -pinene. *Atmospheric Environment. Part A. General Topics*
767 **25**, 997–1008 (1991).
- 768 15. S. Hatakeyama, K. Izumi, T. Fukuyama, H. Akimoto, N. Washida, Reactions of OH with α -
769 pinene and β -pinene in air: Estimate of global CO production from the atmospheric
770 oxidation of terpenes. *Journal of Geophysical Research: Atmospheres* **96**, 947–958 (1991).
- 771 16. S.-H. Zhang, M. Shaw, J. H. Seinfeld, R. C. Flagan, Photochemical aerosol formation from
772 α -pinene- and β -pinene. *Journal of Geophysical Research: Atmospheres* (1992).
- 773 17. T. Hoffmann, J. R. Odum, F. Bowman, D. Collins, D. Klockow, R. C. Flagan, J. H. Seinfeld,
774 Formation of Organic Aerosols from the Oxidation of Biogenic Hydrocarbons. *Journal of*
775 *Atmospheric Chemistry* **26**, 189–222 (1997).
- 776 18. R. J. Griffin, D. R. Cocker III, R. C. Flagan, J. H. Seinfeld, Organic aerosol formation from
777 the oxidation of biogenic hydrocarbons. *Journal of Geophysical Research: Atmospheres*
778 **104**, 3555–3567 (1999).
- 779 19. S. N. Pandis, R. A. Harley, G. R. Cass, J. H. Seinfeld, Secondary organic aerosol formation
780 and transport. *Atmospheric Environment. Part A. General Topics* **26**, 2269–2282 (1992).

- 781 20. S. H. Chung, J. H. Seinfeld, Global distribution and climate forcing of carbonaceous
782 aerosols. *Journal of Geophysical Research: Atmospheres* **107**, AAC 14-1-AAC 14-33
783 (2002).
- 784 21. K. Tsigaridis, M. Kanakidou, Global modelling of secondary organic aerosol in the
785 troposphere: a sensitivity analysis. *Atmospheric Chemistry and Physics* **3**, 1849–1869
786 (2003).
- 787 22. C. Song, K. Na, D. R. Cocker, Impact of the Hydrocarbon to NO_x Ratio on Secondary
788 Organic Aerosol Formation. *Environ. Sci. Technol.* **39**, 3143–3149 (2005).
- 789 23. K. S. Docherty, P. J. Ziemann, Effects of Stabilized Criegee Intermediate and OH Radical
790 Scavengers on Aerosol Formation from Reactions of β-Pinene with O₃. *Aerosol Science and
791 Technology* **37**, 877–891 (2003).
- 792 24. D. Johnson, M. E. Jenkin, K. Wirtz, M. Martin-Reviejo, D. Johnson, M. E. Jenkin, K. Wirtz,
793 M. Martin-Reviejo, Simulating the Formation of Secondary Organic Aerosol from the
794 Photooxidation of Toluene. *Environ. Chem.* **1**, 150–165 (2004).
- 795 25. K. S. Docherty, W. Wu, Y. B. Lim, P. J. Ziemann, Contributions of Organic Peroxides to
796 Secondary Aerosol Formed from Reactions of Monoterpenes with O₃. *Environ. Sci. Technol.*
797 **39**, 4049–4059 (2005).
- 798 26. A. A. Presto, K. E. Huff Hartz, N. M. Donahue, Secondary Organic Aerosol Production
799 from Terpene Ozonolysis. 2. Effect of NO_x Concentration. *Environ. Sci. Technol.* **39**, 7046–
800 7054 (2005).
- 801 27. J. H. Kroll, N. L. Ng, S. M. Murphy, R. C. Flagan, J. H. Seinfeld, Secondary Organic
802 Aerosol Formation from Isoprene Photooxidation. *Environ. Sci. Technol.* **40**, 1869–1877
803 (2006).
- 804 28. M. Martín-Reviejo, K. Wirtz, Is Benzene a Precursor for Secondary Organic Aerosol?
805 *Environ. Sci. Technol.* **39**, 1045–1054 (2005).
- 806 29. J. H. Kroll, N. L. Ng, S. M. Murphy, R. C. Flagan, J. H. Seinfeld, Secondary organic aerosol
807 formation from isoprene photooxidation under high-NO_x conditions. *Geophysical Research
808 Letters* **32** (2005).
- 809 30. N. L. Ng, P. S. Chhabra, A. W. H. Chan, J. D. Surratt, J. H. Kroll, A. J. Kwan, D. C.
810 McCabe, P. O. Wennberg, A. Sorooshian, S. M. Murphy, N. F. Dalleska, R. C. Flagan, J. H.
811 Seinfeld, Effect of NO_x level on secondary organic aerosol (SOA) formation from the
812 photooxidation of terpenes. *Atmospheric Chemistry and Physics* **7**, 5159–5174 (2007).
- 813 31. N. L. Ng, J. H. Kroll, A. W. H. Chan, P. S. Chhabra, R. C. Flagan, J. H. Seinfeld, Secondary
814 organic aerosol formation from *m*-xylene, toluene, and benzene. *Atmospheric Chemistry and
815 Physics* **7**, 3909–3922 (2007).
- 816 32. C. Song, K. Na, B. Warren, Q. Malloy, D. R. Cocker, Secondary Organic Aerosol Formation
817 from *m*-Xylene in the Absence of NO_x. *Environ. Sci. Technol.* **41**, 7409–7416 (2007).

- 818 33. H. O. T. Pye, A. W. H. Chan, M. P. Barkley, J. H. Seinfeld, Global modeling of organic
819 aerosol: The importance of reactive nitrogen (NO_x and NO₃). *Atmospheric Chemistry and*
820 *Physics* **10**, 11261–11276 (2010).
- 821 34. A. A. Presto, N. M. Donahue, Investigation of α -Pinene + Ozone Secondary Organic
822 Aerosol Formation at Low Total Aerosol Mass. *Environ. Sci. Technol.* **40**, 3536–3543
823 (2006).
- 824 35. J. Zhang, K. E. Huff Hartz, S. N. Pandis, N. M. Donahue, Secondary Organic Aerosol
825 Formation from Limonene Ozonolysis: Homogeneous and Heterogeneous Influences as a
826 Function of NO_x. *J. Phys. Chem. A* **110**, 11053–11063 (2006).
- 827 36. D. K. Henze, J. H. Seinfeld, Global secondary organic aerosol from isoprene oxidation.
828 *Geophysical Research Letters* **33** (2006).
- 829 37. D. K. Henze, J. H. Seinfeld, N. L. Ng, J. H. Kroll, T.-M. Fu, D. J. Jacob, C. L. Heald, Global
830 modeling of secondary organic aerosol formation from aromatic hydrocarbons: high- vs.
831 low-yield pathways. *Atmospheric Chemistry and Physics* **8**, 2405–2420 (2008).
- 832 38. C. L. Heald, D. K. Henze, L. W. Horowitz, J. Feddema, J. F. Lamarque, A. Guenther, P. G.
833 Hess, F. Vitt, J. H. Seinfeld, A. H. Goldstein, I. Fung, Predicted change in global secondary
834 organic aerosol concentrations in response to future climate, emissions, and land use change.
835 *Journal of Geophysical Research Atmospheres*, doi: <https://doi.org/10.1029/2007JD009092>
836 (2008).
- 837 39. J. Peeters, J.-F. Muller, HO_x radical regeneration in isoprene oxidation via peroxy radical
838 isomerisations. II : experimental evidence and global impact. *Physical Chemistry Chemical*
839 *Physics* **12**, 14227–14235 (2010).
- 840 40. J. D. Crounse, L. B. Nielsen, S. Jørgensen, H. G. Kjaergaard, P. O. Wennberg, Autoxidation
841 of organic compounds in the atmosphere. *Journal of Physical Chemistry Letters* **4**, 3513–
842 3520 (2013).
- 843 41. M. Ehn, J. A. Thornton, E. Kleist, M. Sipilä, H. Junninen, I. Pullinen, M. Springer, F.
844 Rubach, R. Tillmann, B. Lee, F. Lopez-Hilfiker, S. Andres, I. H. Acir, M. Rissanen, T.
845 Jokinen, S. Schobesberger, J. Kangasluoma, J. Kontkanen, T. Nieminen, T. Kurtén, L. B.
846 Nielsen, S. Jørgensen, H. G. Kjaergaard, M. Canagaratna, M. D. Maso, T. Berndt, T. Petäjä,
847 A. Wahner, V. M. Kerminen, M. Kulmala, D. R. Worsnop, J. Wildt, T. F. Mentel, A large
848 source of low-volatility secondary organic aerosol. *Nature* **506**, 476–479 (2014).
- 849 42. M. P. Rissanen, NO₂ Suppression of Autoxidation–Inhibition of Gas-Phase Highly Oxidized
850 Dimer Product Formation. *ACS Earth Space Chem.* **2**, 1211–1219 (2018).
- 851 43. K. J. Nihill, Q. Ye, F. Majluf, J. E. Krechmer, M. R. Canagaratna, J. H. Kroll, Influence of
852 the NO/NO₂ Ratio on Oxidation Product Distributions under High-NO Conditions. *Environ.*
853 *Sci. Technol.* **55**, 6594–6601 (2021).
- 854 44. A. W. H. Chan, M. N. Chan, J. D. Surratt, P. S. Chhabra, C. L. Loza, J. D. Crounse, L. D.
855 Yee, R. C. Flagan, P. O. Wennberg, J. H. Seinfeld, Role of aldehyde chemistry and NO_x
856 concentrations in secondary organic aerosol formation. *Atmospheric Chemistry and Physics*
857 **10**, 7169–7188 (2010).

- 858 45. M. D. Keywood, J. H. Kroll, V. Varutbangkul, R. Bahreini, R. C. Flagan, J. H. Seinfeld,
859 Secondary Organic Aerosol Formation from Cyclohexene Ozonolysis: Effect of OH
860 Scavenger and the Role of Radical Chemistry. *Environ. Sci. Technol.* **38**, 3343–3350 (2004).
- 861 46. W. Peng, C. Le, W. C. Porter, D. R. I. Cocker, Variability in Aromatic Aerosol Yields under
862 Very Low NO_x Conditions at Different HO₂/RO₂ Regimes. *Environ. Sci. Technol.* **56**, 750–
863 760 (2022).
- 864 47. Y. Baker, S. Kang, H. Wang, R. Wu, J. Xu, A. Zanders, Q. He, T. Hohaus, T. Ziehm, V.
865 Geretti, T. J. Bannan, S. P. O’Meara, A. Voliotis, M. Hallquist, G. McFiggans, S. R. Zorn,
866 A. Wahner, T. Mentel, Impact of HO₂/RO₂ ratio on highly oxygenated α -pinene
867 photooxidation products and secondary organic aerosol formation potential. *EGUsphere*, 1–
868 32 (2023).
- 869 48. T. Berndt, W. Scholz, B. Mentler, L. Fischer, H. Herrmann, M. Kulmala, A. Hansel,
870 Accretion Product Formation from Self- and Cross-Reactions of RO₂ Radicals in the
871 Atmosphere. *Angewandte Chemie International Edition* **57**, 3820–3824 (2018).
- 872 49. Y. Zhao, J. A. Thornton, H. O. T. Pye, Quantitative constraints on autoxidation and dimer
873 formation from direct probing of monoterpene-derived peroxy radical chemistry.
874 *Proceedings of the National Academy of Sciences* **115**, 12142–12147 (2018).
- 875 50. G. McFiggans, T. F. Mentel, J. Wildt, I. Pullinen, S. Kang, E. Kleist, S. Schmitt, M.
876 Springer, R. Tillmann, C. Wu, D. Zhao, M. Hallquist, C. Faxon, M. Le Breton, Å. M.
877 Hallquist, D. Simpson, R. Bergström, M. E. Jenkin, M. Ehn, J. A. Thornton, M. R. Alfarra,
878 T. J. Bannan, C. J. Percival, M. Priestley, D. Topping, A. Kiendler-Scharr, Secondary
879 organic aerosol reduced by mixture of atmospheric vapours. *Nature* **565**, 587–593 (2019).
- 880 51. K. Tsigaridis, N. Daskalakis, M. Kanakidou, P. J. Adams, P. Artaxo, R. Bahadur, Y.
881 Balkanski, S. E. Bauer, N. Bellouin, A. Benedetti, T. Bergman, T. K. Berntsen, J. P. Beukes,
882 H. Bian, K. S. Carslaw, M. Chin, G. Curci, T. Diehl, R. C. Easter, S. J. Ghan, S. L. Gong, A.
883 Hodzic, C. R. Hoyle, T. Iversen, S. Jathar, J. L. Jimenez, J. W. Kaiser, A. Kirkevåg, D.
884 Koch, H. Kokkola, Y. H Lee, G. Lin, X. Liu, G. Luo, X. Ma, G. W. Mann, N. Mihalopoulos,
885 J. J. Morcrette, J. F. Müller, G. Myhre, S. Myriokefalitakis, N. L. Ng, D. O’donnell, J. E.
886 Penner, L. Pozzoli, K. J. Pringle, L. M. Russell, M. Schulz, J. Sciare, Seland, D. T. Shindell,
887 S. Sillman, R. B. Skeie, D. Spracklen, T. Stavrakou, S. D. Steenrod, T. Takemura, P. Tiitta,
888 S. Tilmes, H. Tost, T. Van Noije, P. G. Van Zyl, K. Von Salzen, F. Yu, Z. Wang, Z. Wang,
889 R. A. Zaveri, H. Zhang, K. Zhang, Q. Zhang, X. Zhang, The AeroCom evaluation and
890 intercomparison of organic aerosol in global models. *Atmospheric Chemistry and Physics*
891 **14**, 10845–10895 (2014).
- 892 52. S. J. Pai, C. L. Heald, J. R. Pierce, S. C. Farina, E. A. Marais, J. L. Jimenez, P. Campuzano-
893 Jost, B. A. Nault, A. M. Middlebrook, H. Coe, J. E. Shilling, R. Bahreini, J. H. Dingle, K.
894 Vu, An evaluation of global organic aerosol schemes using airborne observations.
895 *Atmospheric Chemistry and Physics* **20**, 2637–2665 (2020).
- 896 53. W. C. Porter, J. L. Jimenez, K. C. Barsanti, Quantifying Atmospheric Parameter Ranges for
897 Ambient Secondary Organic Aerosol Formation. *ACS Earth Space Chem.* **5**, 2380–2397
898 (2021).
- 899 54. P. O. Wennberg, Let’s Abandon the “High NO_x” and “Low NO_x” Terminology. *ACS EST*
900 *Air*, doi: 10.1021/acsestair.3c00055 (2023).

- 901 55. A. P. Teng, J. D. Crouse, P. O. Wennberg, Isoprene Peroxy Radical Dynamics. *Journal of*
902 *the American Chemical Society* **139**, 5367–5377 (2017).
- 903 56. M. Schervish, N. M. Donahue, Peroxy radical kinetics and new particle formation.
904 *Environmental Science: Atmospheres* **1**, 79–92 (2021).
- 905 57. P. O. Wennberg, K. H. Bates, J. D. Crouse, L. G. Dodson, R. C. McVay, L. A. Mertens, T.
906 B. Nguyen, E. Praske, R. H. Schwantes, M. D. Smarte, J. M. St Clair, A. P. Teng, X. Zhang,
907 J. H. Seinfeld, Gas-Phase Reactions of Isoprene and Its Major Oxidation Products. *Chemical*
908 *Reviews* **118**, 3337–3390 (2018).
- 909 58. E. Praske, R. V. Otkjær, J. D. Crouse, J. C. Hethcox, B. M. Stoltz, H. G. Kjaergaard, P. O.
910 Wennberg, Atmospheric autoxidation is increasingly important in urban and suburban North
911 America. *Proceedings of the National Academy of Sciences of the United States of America*
912 **115**, 64–69 (2018).
- 913 59. L. Xu, K. H. Møller, J. D. Crouse, R. V. Otkjær, H. G. Kjaergaard, P. O. Wennberg,
914 Unimolecular Reactions of Peroxy Radicals Formed in the Oxidation of α -Pinene and β -
915 Pinene by Hydroxyl Radicals. *J. Phys. Chem. A* **123**, 1661–1674 (2019).
- 916 60. E. L. D'Ambro, B. H. Lee, J. Liu, J. E. Shilling, C. J. Gaston, F. D. Lopez-Hilfiker, S.
917 Schobesberger, R. A. Zaveri, C. Mohr, A. Lutz, Z. Zhang, A. Gold, J. D. Surratt, J. C.
918 Rivera-Rios, F. N. Keutsch, J. A. Thornton, Molecular composition and volatility of
919 isoprene photochemical oxidation secondary organic aerosol under low- and high-NO_x
920 conditions. *Atmospheric Chemistry and Physics* **17**, 159–174 (2017).
- 921 61. J. A. Thornton, J. E. Shilling, M. Shrivastava, E. L. D'Ambro, M. A. Zawadowicz, J. Liu, A
922 Near-Explicit Mechanistic Evaluation of Isoprene Photochemical Secondary Organic
923 Aerosol Formation and Evolution: Simulations of Multiple Chamber Experiments with and
924 without Added NO_x. *ACS Earth Space Chem.* **4**, 1161–1181 (2020).
- 925 62. J. E. Shilling, Q. Chen, S. M. King, T. Rosenoern, J. H. Kroll, D. R. Worsnop, K. A.
926 McKinney, S. T. Martin, Particle mass yield in secondary organic aerosol formed by the
927 dark ozonolysis of α -pinene. *Atmospheric Chemistry and Physics* **8**, 2073–2088 (2008).
- 928 63. W. Nie, C. Yan, L. Yang, P. Roldin, Y. Liu, A. L. Vogel, U. Molteni, D. Stolzenburg, H.
929 Finkenzeller, A. Amorim, F. Bianchi, J. Curtius, L. Dada, D. C. Draper, J. Duplissy, A.
930 Hansel, X.-C. He, V. Hofbauer, T. Jokinen, C. Kim, K. Lehtipalo, L. Nichman, R. L.
931 Mauldin, V. Makhmutov, B. Mentler, A. Mizelli-Ojdanic, T. Petäjä, L. L. J. Quéléver, S.
932 Schallhart, M. Simon, C. Tauber, A. Tomé, R. Volkamer, A. C. Wagner, R. Wagner, M.
933 Wang, P. Ye, H. Li, W. Huang, X. Qi, S. Lou, T. Liu, X. Chi, J. Dommen, U. Baltensperger,
934 I. El Haddad, J. Kirkby, D. Worsnop, M. Kulmala, N. M. Donahue, M. Ehn, A. Ding, NO at
935 low concentration can enhance the formation of highly oxygenated biogenic molecules in
936 the atmosphere. *Nat Commun* **14**, 3347 (2023).
- 937 64. A. Lambe, P. Massoli, X. Zhang, M. Canagaratna, J. Nowak, C. Daube, C. Yan, W. Nie, T.
938 Onasch, J. Jayne, C. Kolb, P. Davidovits, D. Worsnop, W. Brune, Controlled nitric oxide
939 production via O(¹D) + N₂O reactions for use in oxidation flow reactor studies. *Atmospheric*
940 *Measurement Techniques* **10**, 2283–2298 (2017).

- 941 65. Z. Peng, J. Lee-Taylor, J. J. Orlando, G. S. Tyndall, J. L. Jimenez, Organic peroxy radical
942 chemistry in oxidation flow reactors and environmental chambers and their atmospheric
943 relevance. *Atmospheric Chemistry and Physics* **19**, 813–834 (2019).
- 944 66. E. Ahlberg, A. Eriksson, W. H. Brune, P. Roldin, B. Svenningsson, Effect of salt seed
945 particle surface area, composition and phase on secondary organic aerosol mass yields in
946 oxidation flow reactors. *Atmospheric Chemistry and Physics* **19**, 2701–2712 (2019).
- 947 67. Z. Li, A. Buchholz, L. M. F. Barreira, A. Ylisirniö, L. Hao, I. Pullinen, S. Schobesberger, A.
948 Virtanen, Isothermal evaporation of α -pinene secondary organic aerosol particles formed
949 under low NO_x and high NO_x conditions. *Atmospheric Chemistry and Physics* **23**, 203–220
950 (2023).
- 951 68. G. M. Wolfe, M. R. Marvin, S. J. Roberts, K. R. Travis, J. Liao, The framework for 0-D
952 atmospheric modeling (F0AM) v3.1. *Geoscientific Model Development* **9**, 3309–3319
953 (2016).
- 954 69. M. E. Jenkin, S. M. Saunders, M. J. Pilling, The tropospheric degradation of volatile organic
955 compounds: a protocol for mechanism development. *Atmospheric Environment* **31**, 81–104
956 (1997).
- 957 70. S. M. Saunders, M. E. Jenkin, R. G. Derwent, M. J. Pilling, Protocol for the development of
958 the Master Chemical Mechanism, MCM v3 (Part A): tropospheric degradation of non-
959 aromatic volatile organic compounds. *Atmospheric Chemistry and Physics* **3**, 161–180
960 (2003).
- 961 71. J. F. Hunter, A. J. Carrasquillo, K. E. Daumit, J. H. Kroll, Secondary Organic Aerosol
962 Formation from Acyclic, Monocyclic, and Polycyclic Alkanes. *Environ. Sci. Technol.* **48**,
963 10227–10234 (2014).
- 964 72. J. A. Thornton, C. Mohr, S. Schobesberger, E. L. D’Ambro, B. H. Lee, F. D. Lopez-Hilfiker,
965 Evaluating Organic Aerosol Sources and Evolution with a Combined Molecular
966 Composition and Volatility Framework Using the Filter Inlet for Gases and Aerosols
967 (FIGAERO). *Accounts of Chemical Research* **53**, 1415–1426 (2020).
- 968 73. R. H. Schwantes, R. C. McVay, X. Zhang, M. M. Coggon, H. Lignell, R. C. Flagan, P. O.
969 Wennberg, J. H. Seinfeld, “Science of the Environmental Chamber” in *Advances in*
970 *Atmospheric Chemistry* (World Scientific, 2017) vol. 1, pp. 1–93.
- 971 74. X. Zhang, J. Ortega, Y. Huang, S. Shertz, G. S. Tyndall, J. J. Orlando, A steady-state
972 continuous flow chamber for the study of daytime and nighttime chemistry under
973 atmospherically relevant NO levels. *Atmos. Chem. Phys* **11**, 2537–2551 (2018).
- 974 75. D. A. Day, J. L. Fry, H. G. Kang, J. E. Krechmer, B. R. Ayres, N. I. Keehan, S. L.
975 Thompson, W. Hu, P. Campuzano-Jost, J. C. Schroder, H. Stark, M. P. DeVault, P. J.
976 Ziemann, K. J. Zarzana, R. J. Wild, W. P. Dubè, S. S. Brown, J. L. Jimenez, Secondary
977 Organic Aerosol Mass Yields from NO₃ Oxidation of α -Pinene and Δ -Carene: Effect of RO₂
978 Radical Fate. *J. Phys. Chem. A* **126**, 7309–7330 (2022).
- 979 76. H. O. T. Pye, B. K. Place, B. N. Murphy, K. M. Seltzer, E. L. D’Ambro, C. Allen, I. R.
980 Piletic, S. Farrell, R. H. Schwantes, M. M. Coggon, E. Saunders, L. Xu, G. Sarwar, W. T.
981 Hutzell, K. M. Foley, G. Pouliot, J. Bash, W. R. Stockwell, Linking gas, particulate, and

- 982 toxic endpoints to air emissions in the Community Regional Atmospheric Chemistry
983 Multiphase Mechanism (CRACMM). *Atmospheric Chemistry and Physics* **23**, 5043–5099
984 (2023).
- 985 77. J. Mao, F. Paulot, D. J. Jacob, R. C. Cohen, J. D. Crouse, P. O. Wennberg, C. A. Keller, R.
986 C. Hudman, M. P. Barkley, L. W. Horowitz, Ozone and organic nitrates over the eastern
987 United States: Sensitivity to isoprene chemistry. *Journal of Geophysical Research:*
988 *Atmospheres* **118**, 11,256–11,268 (2013).
- 989 78. K. R. Travis, D. J. Jacob, J. A. Fisher, P. S. Kim, E. A. Marais, L. Zhu, K. Yu, C. C. Miller,
990 R. M. Yantosca, M. P. Sulprizio, A. M. Thompson, P. O. Wennberg, J. D. Crouse, J. M. St
991 Clair, R. C. Cohen, J. L. Laughner, J. E. Dibb, S. R. Hall, K. Ullmann, G. M. Wolfe, I. B.
992 Pollack, J. Peischl, J. A. Neuman, X. Zhou, Why do models overestimate surface ozone in
993 the Southeast United States? *Atmospheric Chemistry and Physics* **16**, 13561–13577 (2016).
- 994 79. K. H. Bates, D. J. Jacob, A new model mechanism for atmospheric oxidation of isoprene:
995 global effects on oxidants, nitrogen oxides, organic products, and secondary organic aerosol.
996 *Atmospheric Chemistry and Physics* **19**, 9613–9640 (2019).
- 997 80. X. Wang, D. J. Jacob, W. Downs, S. Zhai, L. Zhu, V. Shah, C. D. Holmes, T. Sherwen, B.
998 Alexander, M. J. Evans, S. D. Eastham, J. A. Neuman, P. R. Veres, T. K. Koenig, R.
999 Volkamer, L. G. Huey, T. J. Bannan, C. J. Percival, B. H. Lee, J. A. Thornton, Global
1000 tropospheric halogen (Cl, Br, I) chemistry and its impact on oxidants. *Atmospheric*
1001 *Chemistry and Physics* **21**, 13973–13996 (2021).
- 1002 81. A. B. Guenther, X. Jiang, C. L. Heald, T. Sakulyanontvittaya, T. Duhl, L. K. Emmons, X.
1003 Wang, Model Development The Model of Emissions of Gases and Aerosols from Nature
1004 version 2.1 (MEGAN2. 1): an extended and updated framework for modeling biogenic
1005 emissions. *Geoscientific Model Development* **5**, 1471–1492 (2012).
- 1006 82. E. E. McDuffie, S. J. Smith, P. O'Rourke, K. Tibrewal, C. Venkataraman, E. A. Marais, B.
1007 Zheng, M. Crippa, M. Brauer, R. V. Martin, A global anthropogenic emission inventory of
1008 atmospheric pollutants from sector- and fuel-specific sources (1970–2017): an application of
1009 the Community Emissions Data System (CEDs). *Earth System Science Data* **12**, 3413–3442
1010 (2020).
- 1011 83. J. Krechmer, F. Lopez-Hilfiker, A. Koss, M. Hutterli, C. Stoerner, B. Deming, J. Kimmel,
1012 C. Warneke, R. Holzinger, J. Jayne, D. Worsnop, K. Fuhrer, M. Gonin, J. De Gouw,
1013 Evaluation of a New Reagent-Ion Source and Focusing Ion-Molecule Reactor for Use in
1014 Proton-Transfer-Reaction Mass Spectrometry. *Analytical Chemistry* **90**, 12011–12018
1015 (2018).
- 1016 84. A. Zaytsev, M. Breitenlechner, A. R. Koss, C. Y. Lim, J. C. Rowe, J. H. Kroll, F. N.
1017 Keutsch, Using collision-induced dissociation to constrain sensitivity of ammonia chemical
1018 ionization mass spectrometry (NH₄⁺ CIMS) to oxygenated volatile organic compounds.
1019 *Atmospheric Measurement Techniques* **12**, 1861–1870 (2019).

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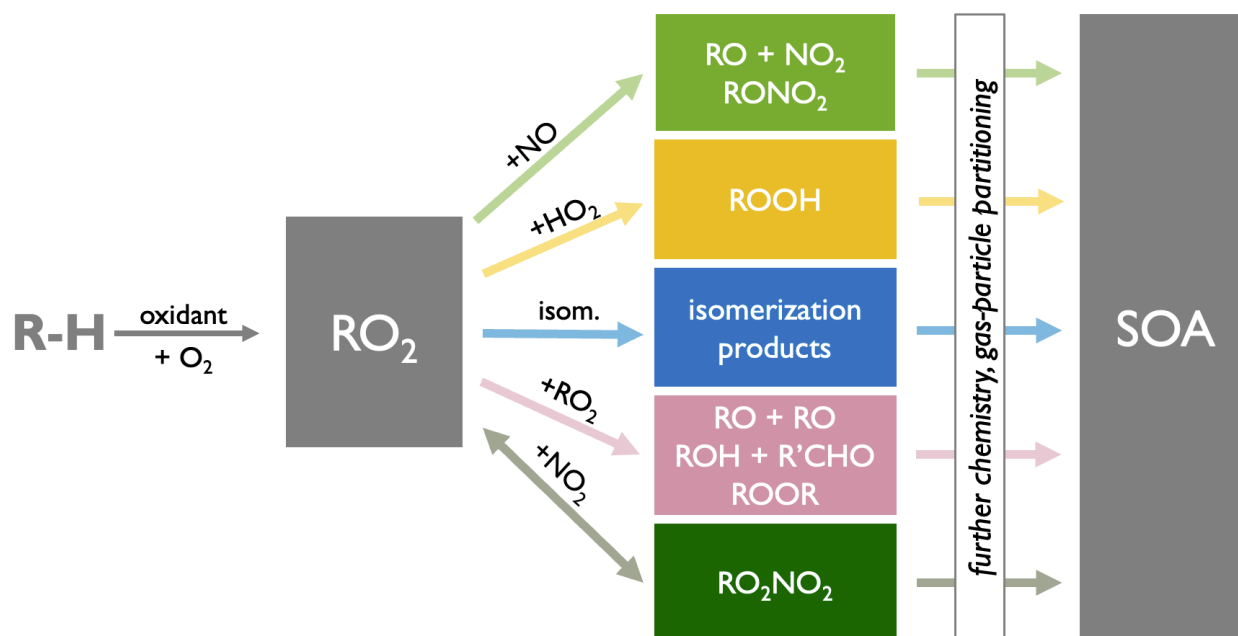
HSK performed the model simulations and created the visualizations. HSK, NT, and MBG performed the experiments. HSK, CLH, and JHK wrote the paper with input from all authors.

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Competing interests: Authors declare that they have no competing interests.

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Data and materials availability: Model and experimental data shown can be found on Zenodo archive XXXXX.



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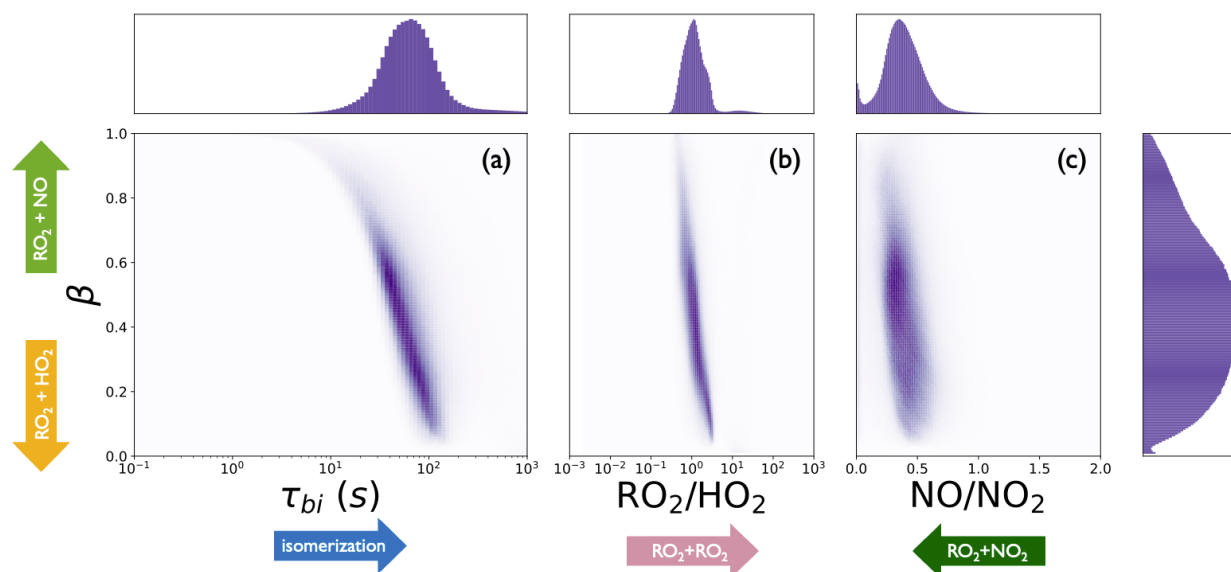
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Fig 1. Schematic of atmospheric RO₂ chemistry and its role in SOA production. RO₂ are formed during VOC oxidation after initial oxidant attach and subsequent molecular oxygen addition. Atmospheric RO₂ can react bimolecularly with NO, HO₂, RO₂, and NO₂, or they can undergo unimolecular isomerization. These RO₂ reaction products can then undergo further chemistry and/or gas-particle partitioning, leading to the formation of SOA.



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Fig. 2. Global atmospheric distribution of RO₂ fates. Shaded regions are 2D histograms of the global distribution of (A) β and τ_{bi} , (B) β and RO₂/HO₂, and (C) β and NO/NO₂, as predicted by GEOS-Chem in hourly output from January and July and weighted by the rate of isoprene + OH oxidation. Surrounding the 2D histograms are projections (1D histograms) of τ_{bi} (a, top), RO₂/HO₂ (b, top), NO/NO₂ (c, top) and β (c, right). β and τ_{bi} are calculated for RO₂ specific to OH-initiated oxidation of isoprene; RO₂/HO₂ is calculated with the sum of all RO₂ radicals (including CH₃O₂).

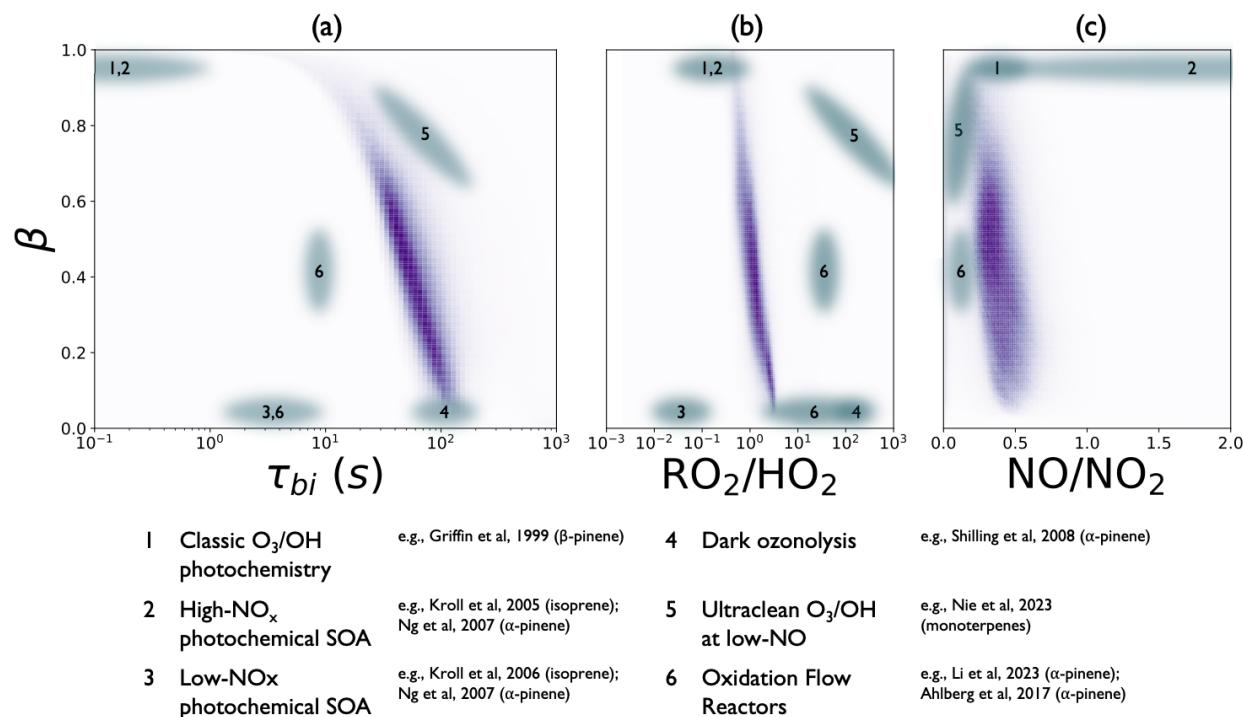
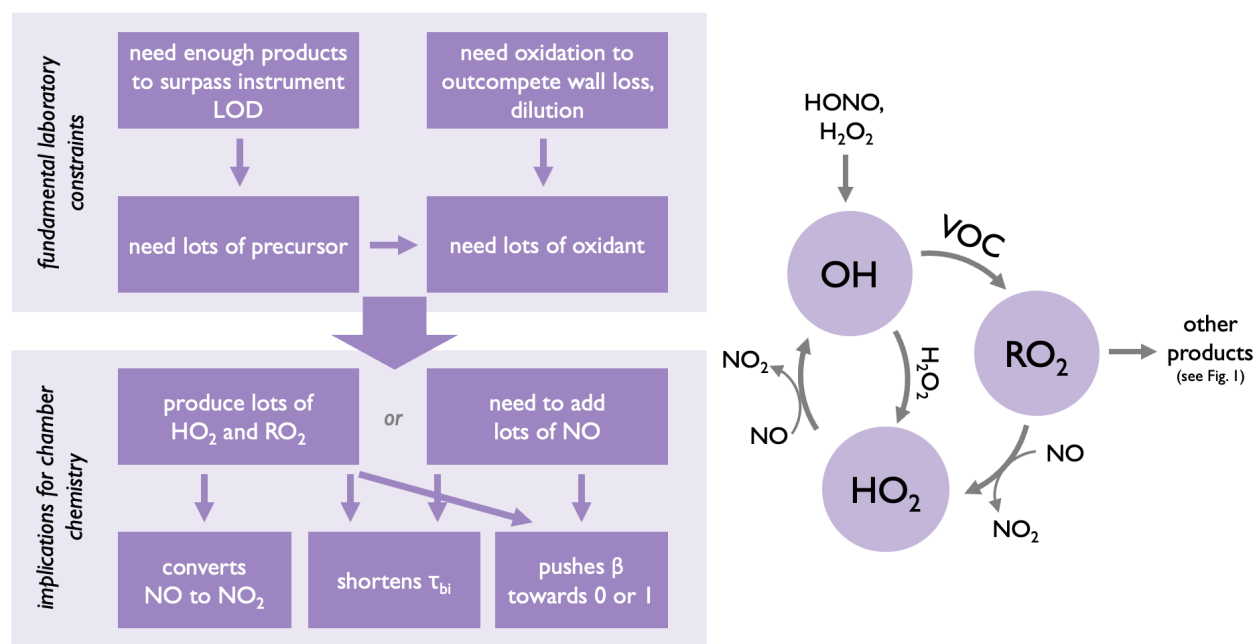
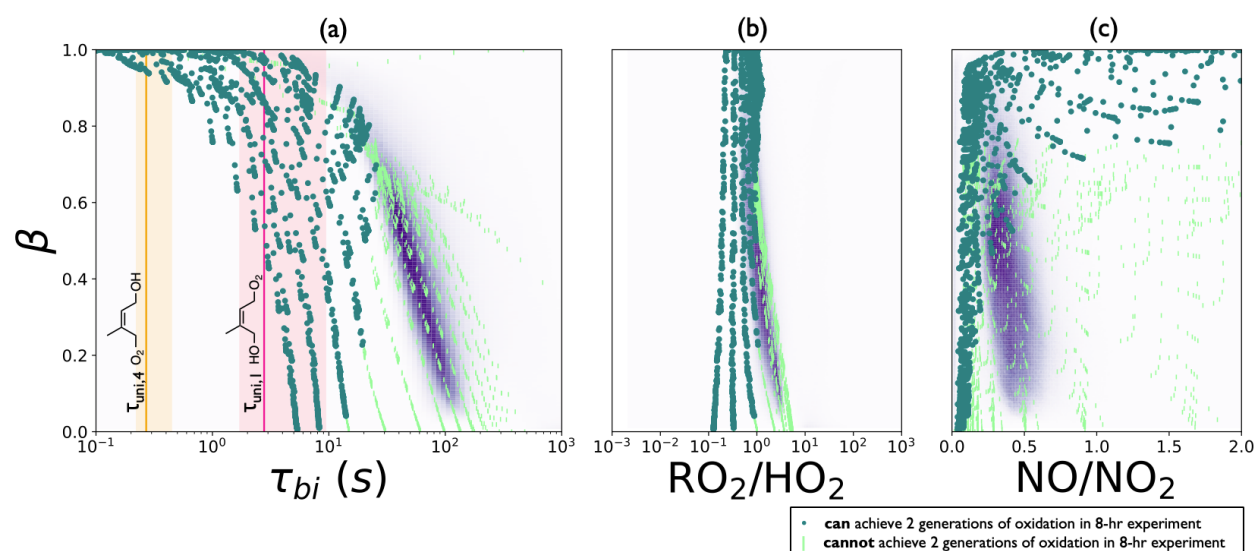


Fig. 3. Comparison of the RO₂ fate distribution achieved during previous approaches to chamber experiments with atmospheric RO₂ fate distributions. Gray ovals show the average values achieved with previous approaches to chamber experiments for (A) β and τ_{bi} , (B) β and RO₂/HO₂, and (C) β and NO/NO₂. In purple are the corresponding 2D histograms showing the global distribution of each parameter as predicted by GEOS-Chem (as in Figure 2). NO/NO₂ ratios are not shown for experiments with $\beta = 0$ (no NO_x). Note that RO₂+RO₂ reactions are included in the τ_{bi} calculation only for experiment type 5 and 6; in all other instances RO₂ + RO₂ reactions contribute <10% to τ_{bi} and are thus omitted for simplicity.



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Fig. 4. Challenges of matching atmospheric conditions in chamber experiments. Left: fundamental constraints of SOA studies and implications for photochemical conditions within chambers. Right: simplified radical chemistry of the VOC oxidation chemistry that underlies such challenges.



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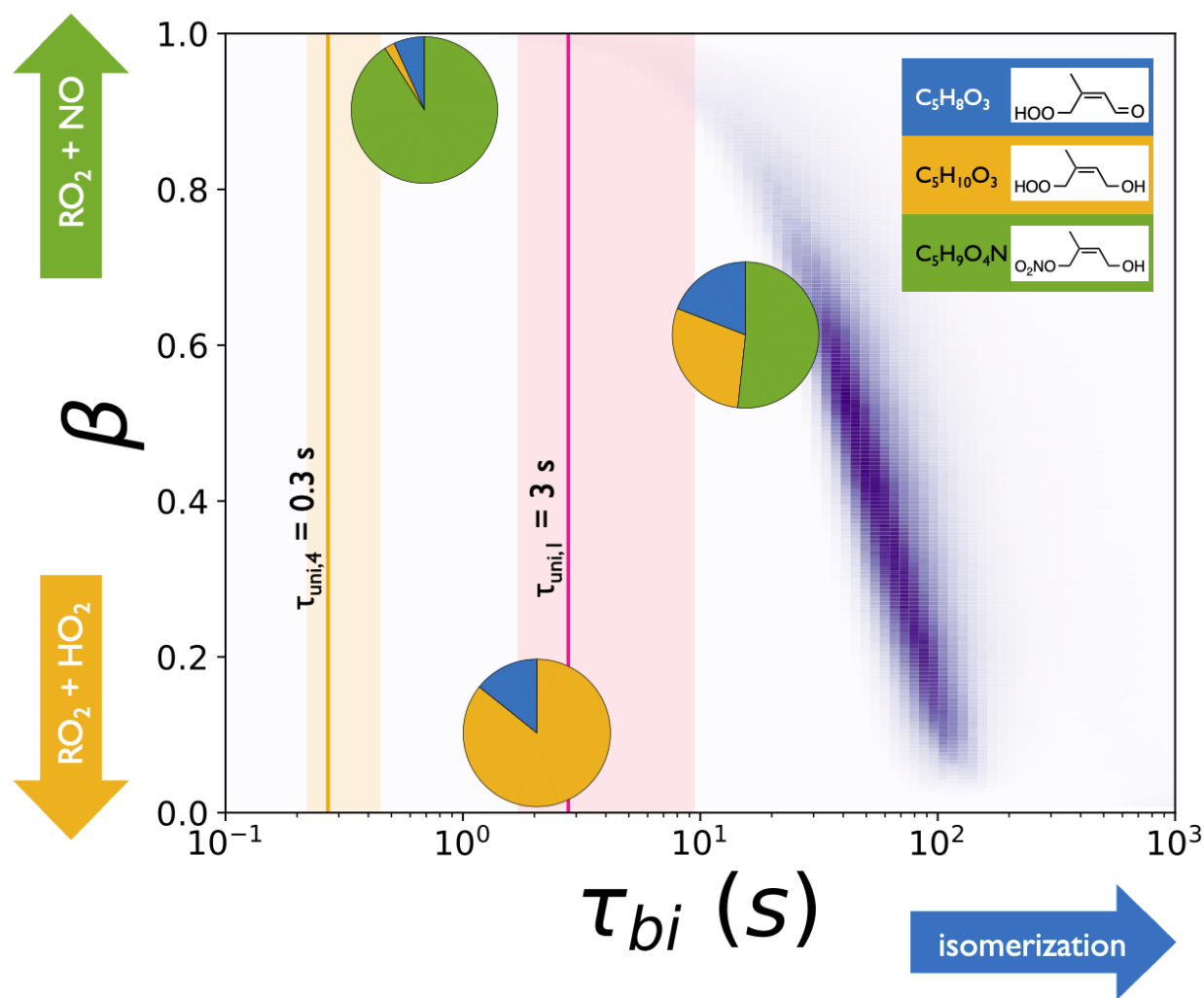
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Fig. 5. Comparison of RO₂ fate distributions achievable in chambers and in the atmosphere.

Points correspond to values of (A) β and τ_{bi} , (B) β and RO_2/HO_2 , and (C) β and NO/NO_2 achievable during theoretical chamber experiments calculated with F0AM initiated with 100 ppb isoprene and different combinations of initial HONO, H₂O₂, and NO. Solid points correspond to 8-hr experiments that achieve second-generation chemistry (defined as final MVK concentration less than half of the maximum MVK concentration); lines correspond to 8-hr experiments that do not achieve sufficient second-generation chemistry. Points are overlaid on 2D histogram of the global distribution of β , τ_{bi} , RO_2/HO_2 , and NO/NO_2 as predicted by GEOS-Chem (as in Figure 2). Note that β and τ_{bi} from GEOS-Chem are calculated with the corresponding atmospheric temperature and pressure whereas the chamber points are all calculated for experiments at 298 K. The unimolecular lifetimes for the two isoprene-derived RO₂ isomers that isomerize in the atmosphere are indicated in pink and orange (Teng et al, 2017). Shaded pink and orange regions indicate the range of unimolecular lifetimes that exist over the 10th – 90th percentile temperature range in the atmosphere during isoprene + OH oxidation.



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Fig. 6. Demonstration of the range of RO₂ fate distributions achievable in chamber experiments of isoprene oxidation. First generation (uncalibrated) gas-phase product yields from isoprene oxidation at a variety of β and τ_{bi} are overlaid on the atmospheric distribution of β and τ_{bi} as shown in Figure 2. The entire product distribution is not represented; rather, three masses measured by NH₄⁺-CIMS representative of the three first-generation product channels are shown. Product yields are calculated as the ratio of product growth (measured by NH₄⁺-CIMS) to the isoprene decay (measured by Vocus PTR) and then normalized to give the relative yields shown in the pie charts here (see Figure S9). Because the data are uncalibrated, fractions shown here should be used only to assess relative differences and are not representative of absolute product distributions. β and τ_{bi} for each experiment are calculated from modeled experiments (F0AM with MCM v3.3.1 chemistry) initialized with measured initial isoprene, HONO, NO, and NO₂ concentrations and with estimated initial H₂O₂ concentrations. Structures shown in the legend are isomers derived from the Z- δ 4-OH, 1-OO isoprene RO₂. The unimolecular lifetimes for the two isoprene-derived RO₂ isomers that can isomerize are indicated in pink and orange (Teng et al, 2017).